




Pirra Jungku and Pirra Warlu: using traditional fire-practice knowledge and contemporary science to guide fire-management goals for desert animals

Sarah Legge^{A,B,C,*} , Hamsini Bijlani^{C,#}, Karajarri Rangers^D, Nurrara Rangers^E, Braedan Taylor^D, Jacqueline Shovellor^D, Frankie McCarthy^E, Chantelle Murray^E, Jesse Ala'i^D, Courtney Brown^D, Kevin Tromp^{E,F}, Sam Bayley^D, Ewan Noakes^{D,G}, Jackie Wemyss^{D,H}, Hannah Cliff^F, Nigel Jackett^G, Bruce Greatwich^G, Ben Corey^G, Mark Cowan^{G,I}, Kristina J. Macdonald^A , Brett P. Murphy^A, Sam Banks^A and Malcom Lindsay^C 

For full list of author affiliations and declarations see end of paper

*Correspondence to:

Sarah Legge
 Research Institute for the Environment and Livihoods, Charles Darwin University, Casuarina, NT 0810, Australia
 Email: sarahmarialegge@gmail.com

#These authors contributed equally to this paper

Handling Editor:

Marlee Hutton

Received: 27 April 2024

Accepted: 15 August 2024

Published: 10 October 2024

Cite this: Legge S *et al.* (2024) Pirra Jungku and Pirra Warlu: using traditional fire-practice knowledge and contemporary science to guide fire-management goals for desert animals. *Wildlife Research* **51**, WR24069. doi:10.1071/WR24069

© 2024 The Author(s) (or their employer(s)). Published by CSIRO Publishing.

This is an open access article distributed under the Creative Commons Attribution-NonCommercial-NoDerivatives 4.0 International License (CC BY-NC-ND).

OPEN ACCESS

ABSTRACT

Context. Indigenous people influenced fire in Australia's deserts for millennia, until colonisation interrupted traditional fire practices. Many groups are reinvigorating those practices to achieve inter-linked biodiversity, cultural, and social benefits. Contemporary fire management integrates modern planning and delivery approaches with cultural process. However, deserts have changed since colonisation (e.g. invasive species, biodiversity loss, climate change) and fire-management outcomes for Country are less certain. Some Indigenous groups are integrating scientific methods into their programs, to examine biodiversity outcomes and refine management. **Aims.** Karajarri and Nurrara Traditional Owners wanted to understand how fire affects Kuwi (mammals, reptiles) on their Country, to guide fire management. Karajarri and Nurrara Rangers worked together on this project, exchanging skills and knowledge with each other, and with scientists. **Methods.** We established 32 monitoring sites at four locations on Karajarri and Nurrara Country. We undertook 11 survey trips from 2018 to 2022, gathering data from 81 site visits. Using Landsat satellite imagery we described fire patterns around each site, focussing on seral stage (as defined by Traditional Owners), patch size, and seral diversity at small (500 m), medium (1 km), and large (5 km) spatial scales. **Results.** There were 378 mammal captures (11 species) and 3392 reptile captures (66 species). Mammal species richness and captures were lowest in Wuntara (recently burnt) and highest in Nyirrinyanu (mature) spinifex. Mammal species richness increased with seral diversity within 1 km around the trapping site. Reptile species richness and captures were unaffected by seral stage. However, many species were caught significantly more often in specific seral stages. Neither reptile richness nor capture rate were related to patch size or seral diversity. **Conclusions.** Retaining a mix of seral stages can maintain diverse reptile and mammal assemblages. The management challenge will be increasing the extent of mature/long-unburnt spinifex, currently the rarest seral stage. Ranger fire management has already reduced mean fire size by 40–52%, and further reductions are possible. **Implications.** Our work improves the understanding of fire effects on northern desert fauna, guides management, provides important ecological information from a little-studied region of Australia's deserts, and increases the recognition of Indigenous conservation management of Country.

Keywords: adaptive management, biodiversity monitoring, Caring for Country, desert fauna, fire ecology, fire management, Indigenous conservation management, Indigenous rangers, species succession.

Introduction

Fire has been part of Australia's ecosystems, including its deserts, through evolutionary time (Martin 2006). Compared with the deserts of other continents, Australian deserts

SHORT SUMMARY AND TRANSLATION

Karajarri and Ngurrara Rangers collaborated with scientists to investigate how fire management in the Great Sandy Desert could help mammals and reptiles. We found that some species prefer particular spinifex post-fire ages, from recently burnt to long-unburnt, and that patchworks of spinifex ages benefit mammals. This means that we should aim to reduce fire size and increase the area of mature/long-unburnt spinifex. Our work provides ecological information from a little-studied region of Australia, and highlights the role of Indigenous land managers in biodiversity conservation.

Translated summary: The two ranger groups have decided not to translate the summary into their languages. Although there are many language speakers from Ngurrara, the most fluent Karajarri speaker passed away last year, and translating complex sentences is now challenging. Out of mutual respect for their collaboration, the rangers prefer not to include a summary that is only in a Ngurrara language. However, we use words from both languages in the text below. A short video about our project can be seen [here](#).

experience fire frequently (Morton *et al.* 2011) for the following four reasons: they experience slightly higher rainfall; the rainfall is strongly seasonal (especially in northern deserts); the vegetation is dominated by a highly flammable grass (spinifex, *Triodia* spp.); and lightning ignitions are common (Edwards *et al.* 2008; Pausas and Keeley 2009; Bradstock 2010). Average fire frequencies depend on annual rainfall and its seasonality, and vegetation type, varying from once every 3–4 years in north-western deserts where rainfall is highest, to once every 20–30+ years in southern deserts.

Following their arrival to the deserts about 50,000 years ago, Aboriginal peoples developed fire practices for many uses, including to manage resources, signal to neighbours, protect sites from wildfire, and ‘clean up’ Country (Gould 1971; Kimber 1983; Lowe 2002; Bird *et al.* 2005; Pike 2008). Cultural protocols determined who burned, and when, according to seasonal calendars and rainfall. These fire practices influenced the fire regime (Bowman 1998) by increasing the number of ignitions, broadening the seasons when fire occurred, decreasing fire size and severity, and increasing fine-scale seral diversity (Burrows and Christensen 1991; Burrows *et al.* 2006; Blackwood *et al.* 2022). People used deserts unevenly through space and time, so some regions may have been burnt intensively by people over some periods, whereas fire was probably applied sparingly in other regions and periods. Thus, the fire regimes at any location were likely to be an interplay between the extent of Indigenous burning and the influence of climate on fire activity (Bliege Bird *et al.* 2012, 2016; Wright *et al.* 2021).

With European colonisation, the Australian deserts were depopulated over a short timeframe. The people-managed

fire regime was replaced (at least in some areas, and times, Wright *et al.* 2021) by a wildfire-dominated system of extensive, severe fires, that regularly exceed 10,000 km² in size (Edwards *et al.* 2008; Turner *et al.* 2011). This change in a key disturbance regime may have contributed to declines of many desert plant and animal species (Allan and Southgate 2002; Greenwood *et al.* 2022). The most profound declines occurred in small to medium-sized mammals; about 60% of species in this group have become extinct since colonisation, and many other species have undergone marked range contractions (Burbidge and McKenzie 1989; McKenzie *et al.* 2007; Woinarski *et al.* 2015). As well as the direct changes to habitat quality wrought by the fire regime shift, the arrival of introduced species, and interactions between fire and invasives, are likely to have led to these declines. For example, frequent, severe fire can facilitate predation by introduced cats (*Felis catus*) and foxes (*Vulpes vulpes*), leading to population decline in native species (Lundie-Jenkins 1993; Short and Turner 1994; Leahy *et al.* 2016; Murphy *et al.* 2018; Doherty *et al.* 2022).

For desert Indigenous people, traditional fire practice is an indivisible component of healthy Country. Traditional fire practice protects tangible cultural and environmental assets, and intangible values such as intergenerational transfer of cultural and ecological knowledge, maintenance of cultural authority, and connection to Country (Weir *et al.* 2011; Cameron 2020). For many Indigenous groups, traditional fire use has been adapted into a contemporary context, incorporating new technologies (e.g. aerial incendiary, remote sensing of burnt areas) and coordinated management programs that typically aim to restore ‘right-way fire’ across traditional lands, for intertwined biodiversity and cultural outcomes (Yanunijarra Aboriginal Corporation 2012; KTLA 2014; Putnis *et al.* 2021; Ruscalleda-Alvarez *et al.* 2023).

Desert Indigenous people have a deep knowledge of how fire affects Country and culturally significant species, including bushmeat species (Lowe 2002; Pike 2008). However, Country has changed through the introduction of new predators (cats, foxes), herbivores (e.g. camels *Camelus dromedarius*) and weeds (e.g. buffel grass, *Cenchrus ciliaris*), and the direct and interacting effects of climate change, which means that traditional practice is adapting to the contemporary context. Moreover, although the new approaches used to complement on-ground burning, such as aerial incendiary, are invaluable for enabling small teams of Indigenous rangers to implement fire management over very large, remote areas (>3000 km²), the resulting fire patterns differ from those arising from traditional fire practices (Perry *et al.* 2018; Ruscalleda-Alvarez *et al.* 2023). Consequently, the impacts of contemporary fire management for biodiversity are less certain.

Context for this study

Scientific knowledge about the effects of fire on much of the Australian desert fauna self-evidently lags behind Indigenous

traditional knowledge, but the scientific method is being increasingly used by Indigenous groups to support their goals for managing fire on Country, in the same way that contemporary operational technologies have been incorporated into traditional fire practice. Indigenous rangers and Traditional Owners are collaborating with scientists to understand and measure how their management affects biocultural assets, and to use this information to adapt management.

In the *Pirra Jungku/Pirra Warlu* project, the Karajarri Rangers initiated a collaboration with scientists from Environs Kimberley, the National Environmental Science Program's Threatened Species Recovery Hub, Western Australia's Department of Biodiversity, Conservation and Attractions, and the Indigenous Desert Alliance, to understand how fire affects desert fauna, so that this information could be used to fine-tune management targets and report on management outcomes to their community and funders. Within a year, the Ngurrara Rangers joined the project. The ranger teams led the project, by setting the research questions and leading the surveys, and organising permits (Karajarri and Ngurrara were the first ranger teams in Western Australia to hold ethics and scientific licences). The project is designed as two-way science, where Indigenous knowledge of Country and scientists' technical expertise are shared to enhance the partnership and understanding.

Karajarri and Ngurrara fire management focusses on achieving the key attributes of 'right-way fire' described by desert peoples: many small fires, creating high seral diversity over given areas, that includes some mature and long-unburnt vegetation (Yanunijarra Aboriginal Corporation 2012; KTLA 2014; IDA 2023; Ruscalleda-Alvarez *et al.* 2023). As part of the *Pirra Jungku* project, we carried out an analysis of historical aerial photographs over Karajarri Country from the 1940s, when people were still living traditional nomadic lifestyles. This work described the fire sizes and patch density created by traditional fire practices; fires were a tenth of the size they are today, seral diversity in a given area was higher, and there was more mature and long-unburnt vegetation, making the average post-fire vegetation age almost 50% older than in 2020 (Blackwood *et al.* 2022). Given these differences, the ranger teams wanted to know what the implications for desert fauna were, and whether they needed to strive for even smaller fires and finer-scale mosaics. 'We want to see what animals are here, check out if the Country is in good shape, and if fire management is working' [Sheen Kitty].

Here, we use data collected over 5 years of monitoring on Karajarri and Ngurrara Country to understand how mammals and reptile communities respond to fire, especially to variation in the age and size of seral stages, and to seral diversity. We focussed on these ground-dwelling groups because we expected them to be most sensitive to changes in ground-layer structure caused by fire. This purpose of the study is to inform key targets for the rangers' fire-management programs, including fire sizes to aim for when conducting prescribed burning, and whether creating fine-scale seral

mosaics and increasing the extent of long-unburnt vegetation is important. The Great Sandy Desert has been little-surveyed by scientists. The only broad-scale sampling within our study area was conducted over 40 years ago (McKenzie 1981), and our information on species present adds to fundamental scientific knowledge about species' distributions.

Methods

Language conventions: we introduce some Karajarri and Ngurrara words in underlined and bold font respectively. Where words from both languages are used, we present them in the order Karajarri/Ngurrara.

Study area

The study was conducted in the Great Sandy Desert of north-western Australia, between 2018 and 2022. The Karajarri Native Title Determination covers 32,000 km² south and inland from Bidadanga and 80 Mile Beach (Fig. 1). Most of the determination (25,000 km²) is designated as an Indigenous Protected Area. Conservation and cultural management is undertaken by Karajarri Rangers, with oversight from elders. The Ngurrara Native Title Determination neighbours Karajarri's eastern boundary and extends south of Fitzroy Crossing (Fig. 1). It covers 76,000 km² and includes the Warlu Jilajaa Jumu Indigenous Protected Area (16,400 km²). The Ngurrara Rangers look after Ngurrara Country, under the direction of their cultural leaders. In this part of Australia, people living a nomadic lifestyle left the deserts late compared with most other regions. The main exodus of the Ngurrara and Karajarri people to coastal missions and pastoral stations occurred between the 1900s and 1960s, with the final departures occurring in the early 1980s (Lowe 2002; Bent *et al.* 2016). 'My grandmother used to travel through here. She told us lots of stories from these places, and she can remember animals that are gone now' [J. Shovellor].

Rainfall in these north-western deserts is influenced by the tropical monsoon. Rainfall is therefore higher, and more seasonal than in more southerly deserts, but remains variable among years. Karajarri, being closer to the coast, receives more rain than Ngurrara. Annual rainfall since 1980 has averaged 502 mm (range 124–1093 mm) for Karajarri, and 403 mm (range 110–698 mm) for Ngurrara (based on interpolated rainfall datasets from SILO (<https://www.longpaddock.qld.gov.au/silo/>), with the points for each selected as midway between the locations sampled in this study). Most rain (>90%) falls between December and March when the tropical monsoon dips furthest to the south (Supplementary material Fig. S1).

We conducted our study at two locations (Yilpi and Kulgara) on Karajarri Country, and two locations (Kuduarra and Kurlku) on Ngurrara Country (Fig. 1). The vegetation at all four locations is Maranguru/**Pirntirri** (but usually called

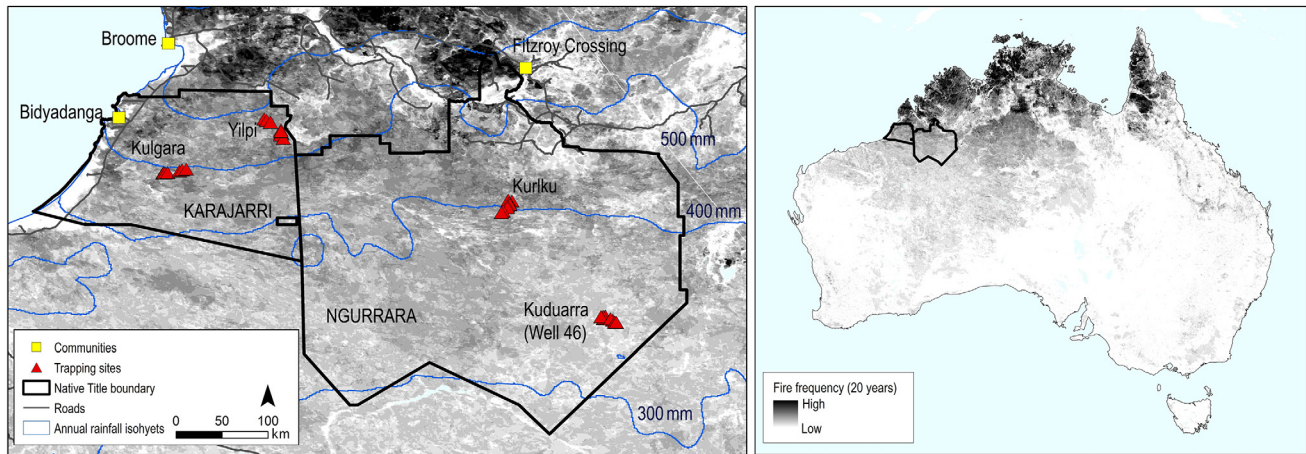


Fig. 1. The Karajarri and Ngurrara Native Title Determinations in the Great Sandy Desert of north-western Australia. The grey shading shows fire frequency over 20 years (2000–2020). The left map shows the four trapping locations, and the annual rainfall isohyets (<http://www.bom.gov.au/climate/maps/averages/rainfall/>).

Jitapuru), spinifex (*Triodia* spp.) sandplains, bordering low Niyamarri/**Jilji**, sand dune country. This is a dominant vegetation group in the Australian deserts (Thackway and Cresswell 1995). Acacias and low trees (4–6 m high) are scattered sparsely, with acacias sometimes forming small, closed canopy thickets that can shade out spinifex (Fig. S2). Vegetative growth in these spinifex deserts is determined by rainfall, with biomass being strongly correlated with the accumulated rainfall since the last fire (Ruscalleda-Alvarez *et al.* 2023). Once spinifex clumps reach sufficient size, fires can travel easily among them. Generally, this occurs every 4–5 years on Karajarri Country, and every 5–8 years on Ngurrara Country, but with a gradient in fire frequency that corresponds to variation in rainfall (Fig. 1).

Fire management

Fire was used intensively by nomadic Karajarri and Ngurrara peoples, to aid with hunting, promote certain food plants (e.g. *Jilarlga*, bush tomato, *Solanum* spp.; *Yukurli*, dodder *Cassytha filiformis*; *Jampar*, bush orange, *Capparis umbonata*; *Mungkarliny*, bush potato, *Ipomoea costata*; *Yarrinyarri*, bush onion, *Cyperus bulbosus*), communicate, ease passage, and ‘clean up’ Country. Maintaining fire practices is considered critical for supporting knowledge transfer, cultural identity, and the health of Country. Burning from the ground is now being supplemented by aerial incendiary techniques, with burning conducted each year in *Marul/Makurra* season (early dry, just after rainy season), to limit the spread of fire in *Laja/Parranga* (hot season), according to a burn plan that is approved by cultural leaders. Both ranger teams have increased the scale and intensity of their fire management since 2019, with the aim of maintaining biodiversity, cultural practice, and the health of Country. ‘*In this desert Country, my grandfather’s Country, we have lots of story places, but it’s been*





hard to get to these areas in the last years to look after things. Now, with helicopters for fire management we can reach places we can’t get to with the vehicle.’ [B. Taylor]. Since landscape-scale, coordinated fire management began in 2019, the mean fire size has been reduced by 52% (to 25.1 km²) on Karajarri Country and 40% (to 21.6 km²) on Ngurrara Country (Karajarri Rangers, Ngurrara Rangers, unpubl. data). The mean fire-return interval across Karajarri and Ngurrara between 2000 and 2023 was 5.2 and 7.1 years respectively (Fig. S1).

Trapping-study design

We set up eight trapping sites to sample *Kuwi*, mammals and reptiles, at each of four locations (i.e. 32 sites in total) (Fig. 1). At each location, sites were positioned 50–500 m off an unsealed access track, stretching over a length of 18–31 km, with sites at least 1 km apart but usually >2 km apart. Sites were established in diverse seral stages, from recently burnt to long-unburnt, and resampled as they aged (Table 1). Eight sites were burned during the study, causing sudden seral-stage changes (five in prescribed burns, three in wildfires). We sampled one to three locations (i.e. 8–24 sites) each year between 2018 and 2022, with each of the four locations being sampled in 2–4 separate years. We undertook surveys at two times of year, namely early dry (April/May), and late dry (September/October), when access is not prevented by rain and temperatures are still high, favouring reptile activity.

At each site, we opened 10 20-L pitfall buckets, four funnel traps (79 cm × 18 cm × 18 cm; Professional Trapping Supplies), and four camera-traps (SWIFT Enduros) set along two parallel 50–60-m-long and 30-cm-high aluminium flywire drift fences set 50 m apart, within a 0.25-ha quadrat. We placed a layer of sand and egg cartons inside pitfalls for shelter, covered with discs of reflective roofing insulation to reduce overheating risk, and give extra protection to small reptiles

Table 1. Spinifex seral stages recognised by Karajarri and Ngurrara people, with the typical range of years since fire this corresponds to, descriptions of the stage, and its ability to carry fire.

English name for seral stage	Karajarri name for seral stage	Ngurrara name for seral stage	
Recently burnt	Jalalu	Wuntara, Nyunma; Waruwaru	
<p>Wuntara: 'after a fire has gone through'</p> <p>Waruwaru: 'after a fire and a little bit of regrowth'</p> <ul style="list-style-type: none"> • 0–1 year since fire • Will not carry fire • 10 site-samples 			
Recovering	Wiyapuka	Parrawa	
<p>'Laparnwanti jitapuru, getting mature spinifex, area where grass has been burnt within the past 3 years or so and spinifex is regrowing but still easy to see turkeys and other bushmeat'</p> <ul style="list-style-type: none"> • 2–3 years on Karajarri; 2–4 years on Ngurrara • Will not carry fire unless wind is very strong • 39 site-samples 			
Mature		Nyirrinyanu	
<p>'Yitilijanka jitapura, grass is dense and mature, stops the hunter from walking and pierces his skin'</p> <ul style="list-style-type: none"> • 4–7 years on Karajarri; 5–9 years on Ngurrara • Will carry fire, but wind is needed at the younger stages • 19 site-samples 			
Long-unburnt		Yurnara	
<p>'Jarlujangka jitapuru, old spinifex from a long time ago, ready for burning again'</p> <ul style="list-style-type: none"> • 8+ years on Karajarri; 10+ years on Ngurrara • Fire spreads easily • 13 site-samples 			

Note the year ranges are slightly different on Ngurrara because it receives less rainfall than does Karajarri. The number of site-samples in each seral stage is shown. In the text, we use the Ngurrara names for seral stages. More photographic examples of habitat and seral stages at sites are in Fig. S2.

from carnivorous mammals. We covered funnels with a 30-cm layer of spinifex, held down with weighted shade-cloth. The camera-traps were strapped to a stake that angled at about 20° off vertical; the cameras were 20–30 cm from the ground

and trained onto the drift fences at about 45° off the perpendicular. They were baited with oats and peanut butter balls to attract small mammals. Surveys were undertaken usually for 4 days, but occasionally up to 7 days, with the survey effort

accounted for in the statistical analysis where appropriate. Traps were checked at dawn, and late afternoon, each day.

Animals captured in pitfall and funnel traps were identified to species. Captured animals were usually released immediately near their point of capture; on occasions when identification was challenging, animals were held for a few hours for other team members to verify identification. All detections of the same species on the same camera-trap in one 12-h (day or night) period were considered as a single capture, to mirror captures in the pitfalls and funnel traps, where an individual can be caught only once in a 12-h period. When small mammals were difficult to identify in camera-trap images, we grouped species. For example, very small rodents, *Jilku/Punypony*, were either sandy inland mice or western delicate mice (*P. hermannsburgensis*, *Pseudomys pilbarensis* sp. nov.) (Roycroft *et al.* 2024); distinguishing these species is challenging even when they are in the hand, so we grouped all captures of small rodents into one 'species type'. Distinguishing between stripe-faced and lesser hairy-footed dunnarts (*Sminthopsis macroura*, *S. crassicaudata*) from images was not possible (thus, grouped to *Sminthopsis* sp.). Sometimes mammals could be identified only as 'native rodent', or even 'mammal'. In these cases, the capture could be used in analyses of overall capture rates but was excluded from species richness analyses when it would artificially inflate the value (e.g. if the only mammal detected at a site was *Sminthopsis* sp., the site had a species richness of one; if the site returned both *S. macroura* and *S. sp.*, the site species richness was still one).

Overall, we conducted 11 survey trips, gathering data from 81 site-surveys (on three surveys, we opened only 4, 6, or 7 sites of the set of 8 for logistical reasons), and ~7200 trap-nights (accounting for some trap closures to manage ants or heat). During 74 site-surveys (i.e. all but seven of the full set), we estimated the characteristics of the ground cover at each site by walking in random lines through the site for about 100 steps and recording whether each toe-fall pointed to bare ground, leaf and woody litter, or grass and herbs, then calculating percentages for these categories.

Trapping trips were guided by two-way science principles. Rangers were trained in, and co-led, scientific aspects of the survey, including understanding the question and management context, the practices of trap-checking, species identification using taxonomic keys and guidebooks, and data entry. Capture totals for each species, in each seral stage, were tallied each day and added to a large scoreboard, so that rangers could see results accumulating in real time. Scientists on the survey learned cultural knowledge where appropriate, including names, stories, and cultural practices such as burning and hunting. A short video describing the trapping surveys is available [here](#).

Fire history

We used Landsat satellite imagery to describe the fine-scale fire patterns around each site. Imagery back to 2008 for

Ngurrara (Landsat 5, 7, 8, 9), and back to 2013 for Karajarri (Landsat 6, 7), was downloaded from the United States Geological Survey (<https://earthexplorer.usgs.gov/>). Fire scars were mapped approximately every 2 months by creating a difference image by using cloud-free images. A manual threshold was set for each image (or each portion of an image if there was a lot of variation within a single difference image) to classify areas of change, i.e. areas that burnt during the two consecutive time periods. We then manually cleaned out false positives owing to cloud cover, water, or other landscape features before exporting the burnt areas. For resourcing reasons, the Landsat series for Karajarri went back only to 2013, which meant that, for sites trapped in 2018, seral stages older than 5 years would be merged. To gain a longer time series we supplemented the Landsat data by downloading processed MODIS imagery for earlier years from the North Australia Fire Service (www.firenorth.org.au/nafi3), so that in every year of trapping, we had a time series of at least 8 years. MODIS imagery has lower spatial resolution than does Landsat (250 vs 30 m). By missing small fire scars, MODIS potentially overestimates fire size and underestimates seral heterogeneity relative to Landsat (Ruscalleda-Alvarez *et al.* 2021). Nevertheless, the value of gaining a longer time series for the sites trapped in earlier years caused us to combine information from the two imagery sources.

Fire metrics

Following Ruscalleda-Alvarez *et al.* (2023), we calculated fire metrics relating to the seral stage (time since fire), patch size, and seral diversity at different scales, for each site, and every year it was sampled. Time since fire was first calculated as the number of wet seasons since the last fire. If a site was sampled in September after a fire burned in May of the same year, the time since fire was zero, but if the fire occurred in May of the previous year, the time since fire was one. These values were then transformed into one of four seral stages that are recognised by Traditional Owners (Table 1), with slight variations between Karajarri and Ngurrara, reflecting the lower rainfall, and thus slower vegetative recovery, of the latter. Seral stages as shown by satellite imagery were validated during surveys at sites. The patch size was the area in hectares of the same-aged vegetation that contained the trapping site. To estimate seral diversity, we calculated the number of vegetation patches of different time-since-fire in buffers around each trapping site of 500 m (small scale), 1 km (medium scale), and 5 km (large-scale) radii. Spatial data manipulation was conducted using ArcMap 10.8.1.

Data preparation and analysis

Fire effects on ground cover

For a subset of site-surveys ($n = 74$), we used ANOVA (in JMP statistical software: https://www.jmp.com/en_au/software.html) to compare the proportion of the ground

that was bare, covered in leaf and woody litter, or covered by grass, across the seral stages.

Immediate effect of fire at sites

We examined the immediate effect of fire on species richness and capture rates, using ANOVA (in JMP) to compare the change in richness and captures (for reptiles and mammals separately) at the eight sites that were burnt between consecutive surveys, versus sites that were sampled in two consecutive surveys without intervening fire.

Species richness and overall abundance with seral succession and diversity

For each survey, and for native mammals and reptiles separately, we calculated the species richness (number of species detected in each class), and the daily capture rate (total captures divided by duration of survey). We did not adjust species richness by survey duration because examination of species accumulation over the survey period showed that 90% of species are captured at least once by the end of the 4th day.

To examine variation in species richness and daily capture rates of mammals and reptiles, we used linear regression mixed models using glmmTMB package v. 1.1.9 (available at <https://doi.org/10.32614/CRAN.package.glmmTMB>) with an underlying Gaussian distribution. We examined model assumptions using DHARMA package v. 0.4.6 (available at <https://doi.org/10.32614/CRAN.package.DHARMA>). To improve the dispersion of residuals in capture-rate models, the square root (mammals) and log (reptiles) of daily capture rates were used as the response variable. Site was fitted as the random term to control for repeat sampling at sites. We examined eight potential fixed effects, including location (Yilpi, Kulgara, Kuduarra, Kurlku), season (early dry, late dry), year (2018, 2019, 2020, 2021, 2022), rainfall in the previous wet season (mm; estimated using interpolated rainfall surface from SILO), seral stage (recently burnt: **Wuntara**; recovering, **Parrawa**; mature: **Nyirrinyanu**; long-unburnt: **Yurnara**), patch size (logged), and seral diversity at the small, medium and large scale (continuous variables). We assessed fixed effects for collinearity. Patch size was correlated ($r > 0.5$) with the seral diversity metrics (especially small and medium scale), and the seral diversity metrics of each scale were correlated with each other (especially small and medium scale), so these variables were not included within the same model. We built candidate models using forward selection of the non-correlated fixed effects and the interaction between location and season. Models were evaluated using the Akaike information criterion corrected for small samples (AICc); best-ranked models were those with the smallest AICc and others with a Δ AICc within two of the best model. When two or more models ranked highly, we undertook model averaging using MumIn package v. 1.47.5 (available at <https://doi.org/10.32614/CRAN.package.MuMIn>). Using emmeans package v. 1.10.1 (available at <https://doi.org/10.32614/CRAN.package.emmeans>) we undertook post-model tests to (1) determine the overall

importance of fixed effects and (2) to examine pairwise comparisons that adjust P -values using a Tukey method to reduce the probability of making a Type I error. We created figures using Sigmplot and ggplot2 package v. 3.5.0 (<https://doi.org/10.32614/CRAN.package.ggplot2>) and analysis was undertaken using R 4.3.3 (R Core Team 2024).

Individual species responses to seral succession and diversity

Using the same analytical approach as for species richness and total capture rates, we explored the responses of individual species to seral succession and diversity. For mammals we investigated the presence/absence of species caught at more than 30% of site-surveys by using a binomial link function. For reptiles, where a larger number of species were caught at a larger number of site-surveys, we were able to fit models to explore variation in abundance. For species that were sufficiently frequently encountered (caught at >50% of the site-surveys), we assessed daily capture rates adjusted with a square-root transformation where needed, to improve the dispersion of residuals. For one species, we used a negative binomial distribution by using daily counts, with an offset of the logged survey duration to improve the dispersion of residuals.

Many Kuwi, mammal and reptile species were caught at less than 30 and 50% of sites, making the analyses described above not possible. We therefore extended our exploration of the seral preferences of species (caught at least eight times) by summing their captures in each of the four seral stages and using chi-squared heterogeneity tests (in JMP) to compare the observed captures against the expected (their total captures split between seral stages according to the proportions in which they were sampled). Correct identification to species is difficult for *Proablephurus reginae/tenuis*, and *Gehyra kimberlyensis/purpurescens*; the species in each case have similar ecologies, so we grouped these captures for analysis. We caught several species of delma (*Delma borea/butleri/desmosa/nasuta/tincta*) each with a low capture total, and we grouped these for analysis.

We also considered whether rare, or rarely detected species (i.e. those caught <6 times across all 81 surveys) were more likely to occur in a seral stage, by testing whether the distribution of summed captures across the stages conformed to expectation, by using a chi-squared heterogeneity test.

Results

Fire effects on ground cover

Ground characteristics varied across seral stages. The proportion of bare ground was highest in **Wuntara** (recently burnt, 62%), and lowest in **Yurnara** (long-unburnt) ($F_{3,70} = 9.82$, $P < 0.0001$; Fig. 2). The proportion of the ground covered by grass or herbs was lowest in **Wuntara** (recently burnt), higher in **Parrawa** (recovering), **Nyirrinyanu** (mature), and **Yurnara** (long-unburnt) ($F_{3,70} = 4.64$, $P = 0.005$;

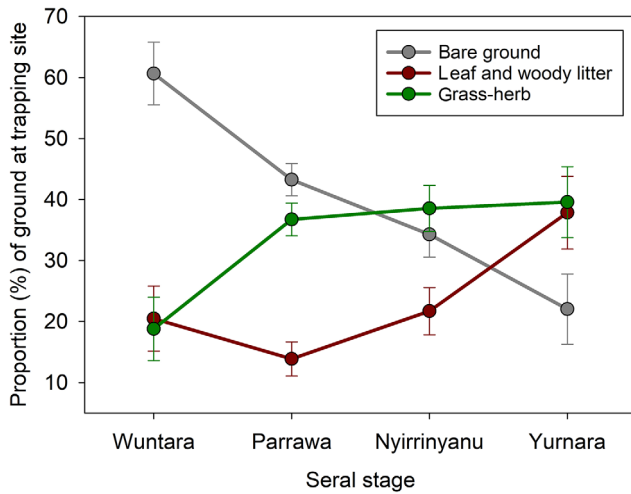


Fig. 2. Changes in ground cover (**Walyara**, bare ground; leafy and woody litter; and grass-herb, **Maramarr**, **Marranguru/Jitapura**) as vegetation recovers after fire from recently burnt **Wuntara**, to recovering **Parrawa**, to mature **Nyirrinyanu**, then to long-unburnt **Yurnara**. Graphs present means with standard errors for site-surveys in each seral stage. Sample for each seral stage = 10, 37, 19, 8 respectively.

Fig. 2. Leaf litter was lowest during **Wuntara** and **Parrawa**, and highest in **Yurnara** ($F_{3,70} = 3.89, P = 0.01$; **Fig. 2**).

Mammal and reptile captures

We recorded 378 mammal captures, and 3392 reptile captures, over the study (5.3 and 47% trapping success respectively). Most mammal captures were on camera-traps (56%), followed by pitfalls (44%). A single mammal was caught in a funnel-trap. Most reptile captures were in pitfalls (82%), and funnel-traps (13%), with 4% of captures on camera-traps. We caught 11 different mammal species (Supplementary material Table S1a). The five species captured most often were *Pseudomys hermannsburgensis/pilbarensis* sp. nov., *Notomys alexis*, *Sminthopsis youngsoni*, *Leggadina lakedownensis*, and *Pseudomys desertor*. Two introduced species, feral cats **Minyawwi/Minyawwi** and European foxes **Ngatin/Kirlmarrara** were also caught on camera-traps (nine captures each, where a capture could be one to multiple detections in a 12-h period), but these captures were omitted from analyses. We caught 65 different reptile species (Table S1b); the five reptiles captured most were *Lerista bipes*, **Pampirra/Wiji** *Ctenophorus isolepis*, **Pantily/Wakura** *Diplodactylus laevis*, **Liwirungu/Wurrkarn** *Ctenotus pantherinus* and **Pantily/Wakura** *Lucasium stenodactylum*.

Immediate effect of fire at sites

Changes in reptile species richness and daily capture rates at sites that burnt between consecutive surveys were not significantly different from the changes at sites that did not burn between consecutive surveys, and none of the changes were

significantly different from zero (richness: $F_{1,45} = 1.59, P = 0.29$; captures: $F_{1,45} = 0.26, P = 0.61$; **Fig. 3**). Sites that burnt experienced a significant reduction in mammal daily capture rates ($F_{1,45} = 5.70, P = 0.021$) but not species richness ($F_{1,45} = 0.37, P = 0.55$; **Fig. 3**).

Species richness and overall abundance with seral succession and diversity

Mammals

Overall species richness ranged from 0 to 4 across all the site-surveys. The best models for mammal species richness included season, seral stage, rainfall, and seral diversity (Table 2a). Species richness significantly increased with rainfall, was highest in **Parrawa** and **Nyirrinyanu** (recovering and mature) seral stages, and increased with seral diversity at the medium scale (1 km) (Table 3a, Fig. 4). Species richness tended to increase in the early dry, but the effect was not significant (Table 3a).

Daily capture rates ranged from 0 to 3.5 mammals/day. The best models for daily capture rate included seral stage, season, and rainfall (Table 2a). Capture rates were significantly higher in **Parrawa** and **Nyirrinyanu** (recovering and mature) seral stages (Fig. 4), whereas seasonal and rainfall differences were not significant (Table 3a).

Reptiles

Overall species richness ranged from 5 to 20 across all the site-surveys. The best models contained location, season (and their interaction), rainfall and patch size (Table 2b). Species richness varied significantly among locations and seasons. At Kulgara and Yilpi, the species richness was higher in the early dry than the late dry, whereas at Kuduarra and Kurlku (the two drier locations, Fig. 1), species richness was similar or increased between the early and late dry seasons (Fig. 5). Rainfall and patch size were in the best-ranked models for

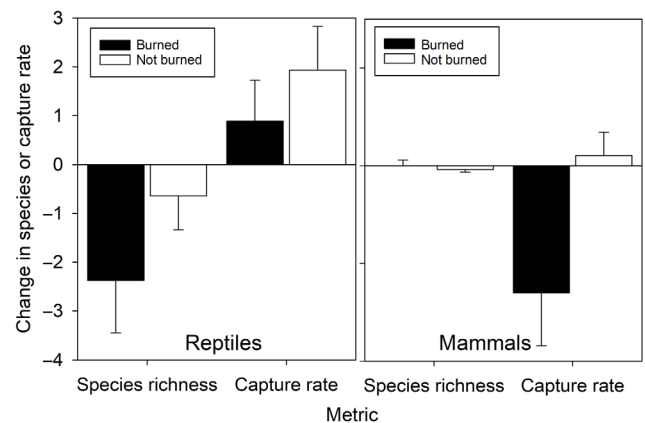


Fig. 3. Changes in species richness and daily capture rates between consecutive surveys at sites that burnt in the interim, or not. Results for reptiles and mammals are shown separately. Bars present means with standard errors.

Table 2. The best-ranked models for each analysis, showing the AICc.

Response variable	Best ranked models	AICc
(a) Mammals		
Species richness	Season + seral_stage + rainfall + PD_1km + (I site)	199.6
Daily capture rate	Seral_stage + (I site)	109.00
	Seral_stage + season + (I site)	110.03
	Seral_stage + rainfall + (I site)	108.21
(b) Reptiles		
Species richness	Location × season + (I site)	396.24
	Location × season + rainfall + (I site)	397.54
	Location × season + log(patch_size)	367.12
	Location × season + rainfall + log(patch_size) + (I site)	398.08
Daily capture rate	Location × season + rainfall + PD_500m + (I site)	44.50
(c) Individual species		
<i>Pseudomys hermannsburgensis/pilbarensis</i> sp. nov. <u>Jilku/Punypuny</u>	PD_500m + (I site)	107.56
	PD_500m + season × location + (I site)	105.88
	PD_500m + season × location + seral_stage + (I site)	107.09
	PD_500m + season × location + seral_stage + rainfall + (I site)	107.09
<i>Notomys alexis</i> <u>Jilku/Punypuny</u>	Location + (I site)	73.27
	Location + rainfall + (I site)	74.99
	Location + PD_1km + (I site)	73.44
	Location + PD_1km + rainfall + (I site)	75.22
<i>Lucasium stenodactylum</i> <u>Pantily/Wakura</u>	Seral_stage + rainfall + (I site)	81.30
	Seral_stage + rainfall + log(patch_size) + (I site)	81.17
	Seral_stage + rainfall + log(patch_size) + season + (I site)	82.95
<i>Ctenopus pantherinus</i> <u>Liwirungu/Wurrkarn</u>	Location + year + (I site)	102.82
	Location + year + seral_stage + (I site)	102.94
<i>Diplodactylus laevis</i> <u>Pantily/Wakura</u>	Location × season + rainfall + (I site)	341.47
	Location × season + rainfall + seral_stage + (I site)	341.91
	Location × season + rainfall + log(patch_size) + (I site)	341.29
	Location × season + rainfall + seral_stage + log(patch_size) + (I site)	343.03
<i>Ctenophorus isolepis</i> <u>Pampirta/Wiji</u>	Location + year + seral_stage + season + (I site)	122.18
<i>Lerista bipes</i> <u>Winkujartu 'lice-eating snake'</u>	Location × season + PD_500m + rainfall + (I site)	117.67
	Location × season + PD_500m + rainfall + seral_stage + (I site)	118.40

The model sets include the best model, and any other models within 2 AICc.

species richness (Table 2b), but neither were significant (Table 3c).

Daily capture rates ranged from 3.6 to 24.5 reptiles/day. In the best model (Table 2b), daily capture rates significantly declined between the early and late dry at Kurlku and Kulgara, but did not change at Yilpi and Kuduarra (Fig. 5). Daily capture rates significantly declined as the seral diversity at the small scale (within 500 m) increased (Table 3b, Fig. 5). Daily capture rates also significantly declined as rainfall decreased, probably driven by patterns in three frequently caught species (see below).

Fire response of individual species

Mammals

Two mammal species were captured at more than 30% of sites. The model results are shown in Tables 2c and 3c. *Pseudomys hermannsburgensis/pilbarensis* sp. nov. occurrence had a significant negative relationship with seral diversity at the small scale (500 m). It was also more likely to be caught in the recovering seral stage, and there were interactions between location and season, but although these effects were in the best models (Table 2c), they were not significant

Table 3. The statistics for each fixed effect in the top-ranked models for each analysis, after model averaging. Significant effects are bolded.

Response variable	Fixed effect	Df1	Df2	F ratio	P-value
(a) Mammals					
Species richness	Season	1	72	3.56	0.0633
	Seral_stage	3	72	9.09	<0.0001
	Rainfall	1	72	5.57	0.0210
	PD_500m	1	72	5.00	0.0285
Daily capture rate	Seral_stage	3	74	7.00	0.0003
	Season	1	74	0.12	0.7248
	Rainfall	1	74	1.00	0.3206
(b) Reptiles					
Species richness	Location	3	69	23.51	<0.0001
	Patch size	1	69	0.37	0.5454
	Rainfall	1	69	0.28	0.5981
	Season	1	69	3.07	0.0842
	Location: season	3	69	6.81	0.0004
	Daily capture rate	Location	3	69	24.50
	Season	1	69	22.08	<0.0001
	Rainfall	1	69	4.97	0.0291
	PD_500m	1	69	5.21	0.0256
	Location: season	3	69	12.81	<0.0001
(c) Individual species					
<i>Pseudomys hermannsburgensis/pilbarensis</i> sp. nov. <u>Jilku/Punypuny</u>	Seral_stage	3	67	0.13	0.9434
	Location	3	67	0.58	0.6315
	PD_500m	1	67	4.02	0.0489
	Rainfall	1	67	0.01	0.9267
	Season	1	67	0.25	0.6162
	Location: Season	3	67	0.84	0.4781
<i>Notomys alexis</i> <u>Jilku/Punypuny</u>	Location	3	74	4.97	0.0034
	PD_1km	1	74	0.47	0.4948
	rainfall	1	74	0.13	0.7215
<i>Lucasium stenodactylum</i> <u>Pantily/Wakura</u>	Seral_stage	3	72	3.55	0.0187
	Patch size	1	72	0.78	0.3794
	Rainfall	1	72	4.17	0.0448
	Season	1	72	0.08	0.7748
<i>Ctenopus pantherinus</i> <u>Liwirungu/Wurrkarn</u>	Seral_stage	3	68	0.26	0.8567
	Location	3	68	3.75	0.0148
	Year	4	68	2.09	0.0923
<i>Diplodactylus laevis</i> <u>Pantily/Wakura</u>	Seral_stage	3	66	0.17	0.9186
	Location	3	66	3.63	0.0173
	Patch size	1	66	0.50	0.4830
	Rainfall	1	66	8.41	0.0051

(Continued on next column)

Table 3. (Continued).

Response variable	Fixed effect	Df1	Df2	F ratio	P-value
	Season	1	66	0.04	0.8486
	Location: season	3	66	7.22	0.003
<i>Ctenophorus isolepis</i> <u>Pampirta/Wiji</u>	Location	3	67	3.03	0.0354
	Year	4	67	9.31	<0.0001
	Seral_stage	3	67	5.68	0.0016
	Season	1	67	8.87	0.0040
<i>Lerista bipes</i> <u>Winkujartu 'lice-eating snake'</u>	Seral_stage	3	66	0.19	0.9006
	Location	3	66	20.14	<0.0001
	PD_500m	1	66	15.58	0.0002
	Rainfall	1	66	20.42	<0.0001
	Season	1	66	18.64	0.0001
	Location: season	3	66	8.47	0.0001

(Table 3c). The occurrence of *Notomys alexis* differed significantly between locations, with the highest probability of occurring at Kurlku. It also increased with seral diversity at the medium scale (1 km), and rainfall, but these effects were not significant (Table 3c, Fig. 6).

Reptiles

Five reptile species were caught at >50% of the site-surveys. Across this set of species, rainfall had a significant negative effect on captures for three of these species, namely *Lerista bipes*, *Diplodactylus laevis*, and *Lucasium stenodactylum* (Tables 2c, 3c, Fig. 6), which probably also drove the lower capture rates for reptiles as rainfall increased (reported above). Higher rainfall drives higher vegetation growth, which may disadvantage these three species, because they were most frequently caught in recently burnt and recovering spinifex (see below). Location was important for four species, but for *Lerista bipes* this depended on season. Season was also a significant factor for *Ctenophorus isolepis*, whose capture rates were higher in the early dry season. Season was included in the best models for *Diplodactylus laevis* and *Lucasium stenodactylum*, but was not significant in both cases. Seral stage was included in top models for all five species, but was a significant predictor for only one species, *Ctenophorus isolepis*, with capture rates highest in the recovering stage. Seral stage diversity at small scale (500 m) was negatively associated with the capture rates of *Lerista bipes* (Tables 2c, 3c, Fig. 6).

Seral preferences of species

Notomys alexis was caught disproportionately often in **Parrawa**, the recovering seral stage. The small *Pseudomys hermannsburgensis/pilbarensis* sp. nov. was caught less than

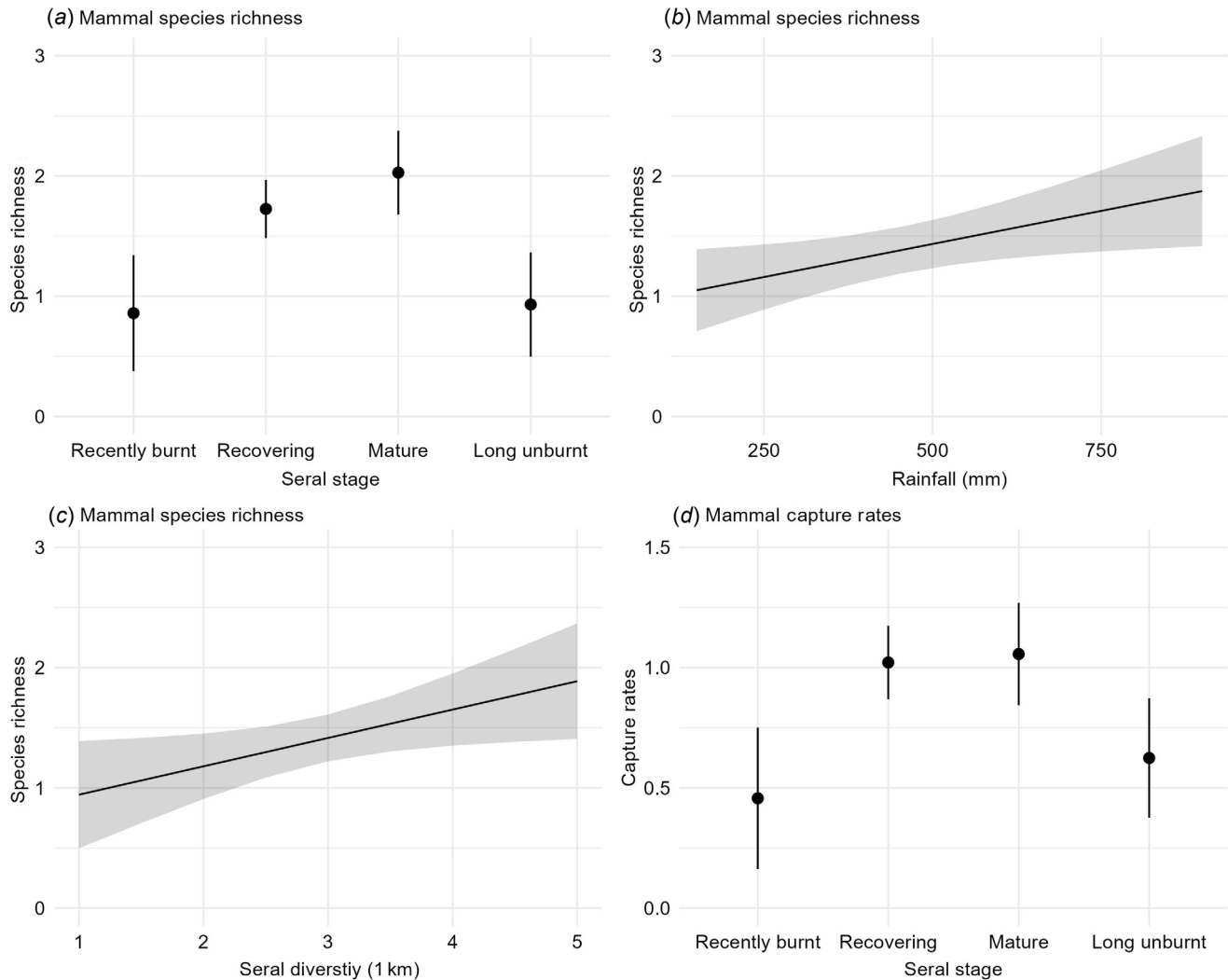


Fig. 4. Significant variables in the best models for mammal species richness and daily capture rates. Values present estimated means and 95% confidence intervals.

expected at **Wuntara**, recently burnt sites, but the larger *Pseudomys desertor* was caught more than expected in **Nyirriyanu**, mature sites. *Sminthopsis youngsoni* captures were spread among seral stages as expected, but when captures of all *Sminthopsis* spp. were considered, there were more than expected in **Nyirriyanu** and fewer in **Wuntara** (Fig. 7; full results in Table S1a).

Three reptile species (*Ctenotus nuchalis*, *Diplodactylus laevis*, *Rhynchoedura ornata*) were caught in **Wuntara**, recently burnt sites, significantly more than expected. Four species (*Ctenophorus slateri*, *Proablephurus reginae*, *Proablephurus tenuis*, *Ctenotus inornatus*) were caught in **Yurnara**, long-unburnt sites, significantly more than expected. Sixteen reptile species were caught significantly more often than expected in the intermediate and older seral stages, especially **Parrawa**, recovering, but also in **Nyirriyanu**, mature, or **Nyirriyanu** and **Yurnara**, mature and long-unburnt. A further 21 species showed no statistical preference for seral stages; some had a

high number of captures, but some had low overall captures, and a greater sample may reveal seral preferences (Fig. 7; full results in Table S1b).

We also considered the distribution of captures of rare species (i.e. caught <6 times) across the seral stages. For reptiles, captures in the four seral stages were 8, 28, 19, and 14, which was not significantly different from expected (9, 33, 16, 11) ($X^2_3 = 3.92$, $P = 0.27$). For mammals, captures were 1, 5, 13, and 2, which was tending to differ from expected (3, 10, 5, 3) by skewing to the **Nyirriyanu**, mature seral stage ($X^2_3 = 6.63$, $P = 0.093$).

Discussion

The importance of traditional fire practices for biodiversity, cultural and social benefits is beginning to be recognised globally (Hoffman *et al.* 2021). Our work contributes to this

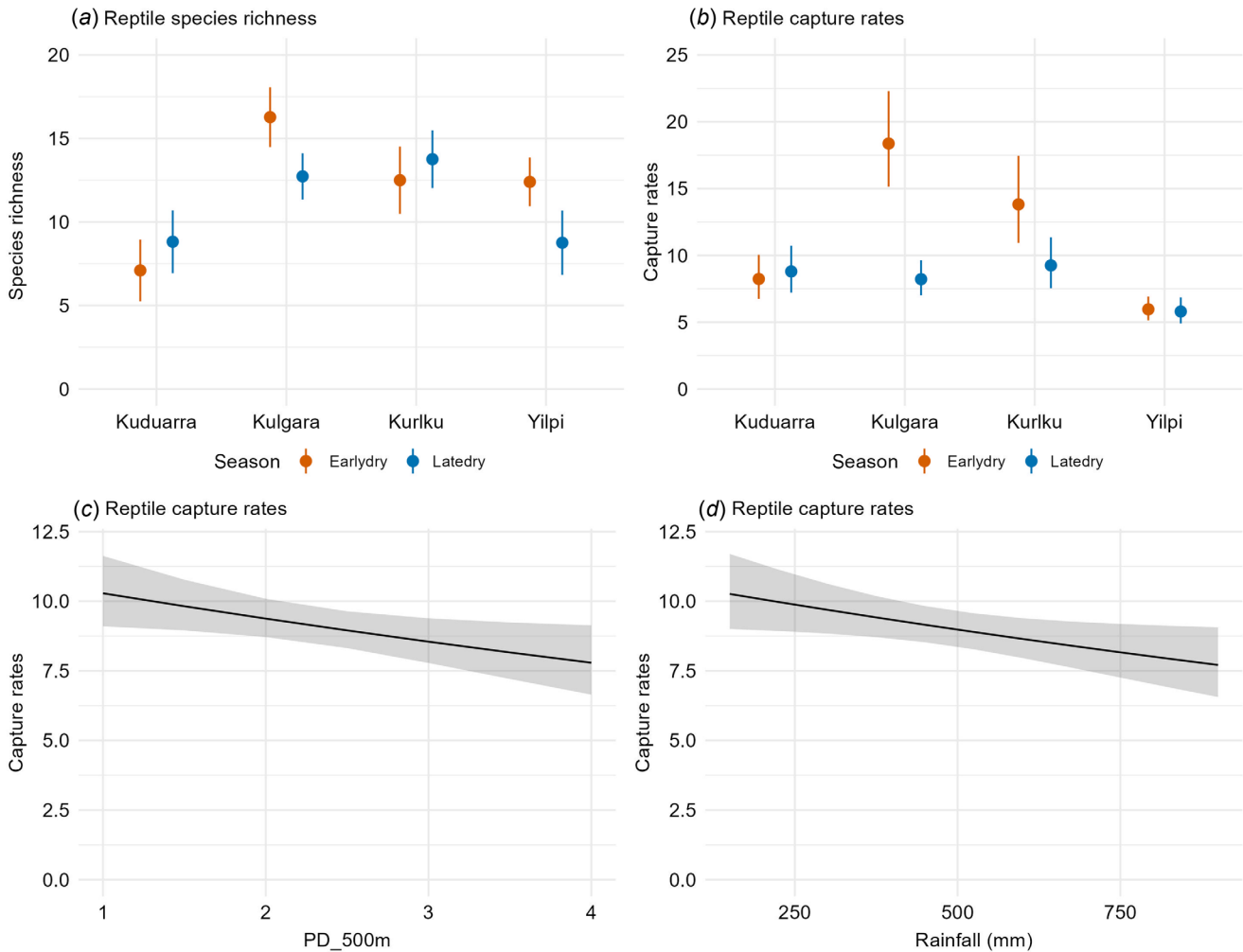


Fig. 5. Significant variables in the best models for reptile species richness and daily capture rates. Figures present estimated means and 95% confidence intervals.

awareness, and provides important ecological information from a little-studied region of Australia's deserts (McKenzie 1981). Desert fire management by Indigenous rangers focusses on creating key attributes of right-way fire, namely, burning many small fires that contribute towards increasing the seral diversity, and the extent of mature and long-unburnt vegetation (IDA 2023; Ruscalleda-Alvarez *et al.* 2023). Our study aimed to understand the benefits of a right-way fire regime for small, ground-dwelling Kuwi, reptiles and mammals, including what size of fire and scale of seral diversity the ranger teams should aim for, and how important it is to protect long-unburnt vegetation. Our results suggest that these reptile and mammal communities respond differently to fire and have different patterns of post-fire recovery. Mammal species richness and capture rates were lowest after fire, recovered as the vegetation recovered, and species richness was positively associated with seral diversity. In contrast, the overall species richness and abundance of reptiles was little-affected by fire or the seral stage, but this result masked strong

associations between many individual species and certain seral stages, including mature and long-unburnt vegetation. Surprisingly, given the strong evidence for species succession with post-fire vegetation recovery, reptile species richness and capture rates were either unrelated, or were negatively related, to increased seral diversity. Reptile responses were also influenced by location, season, and rainfall, whereas mammals responded only to rainfall. Below, we discuss some of the key findings, including significant species detections, and responses to seral stage and size, seral diversity, and long-unburnt vegetation. We finish by discussing what our results mean for management.

Significant species detections

Our surveys showed some significant records. We detected Larnkurr/Jumpiyinti, the brushtail possum, *Trichosurus vulpecula*, at Yilpi, on the edge of the Edgar Ranges on Karajarri Country. Larnkurr was previously detected, in hair

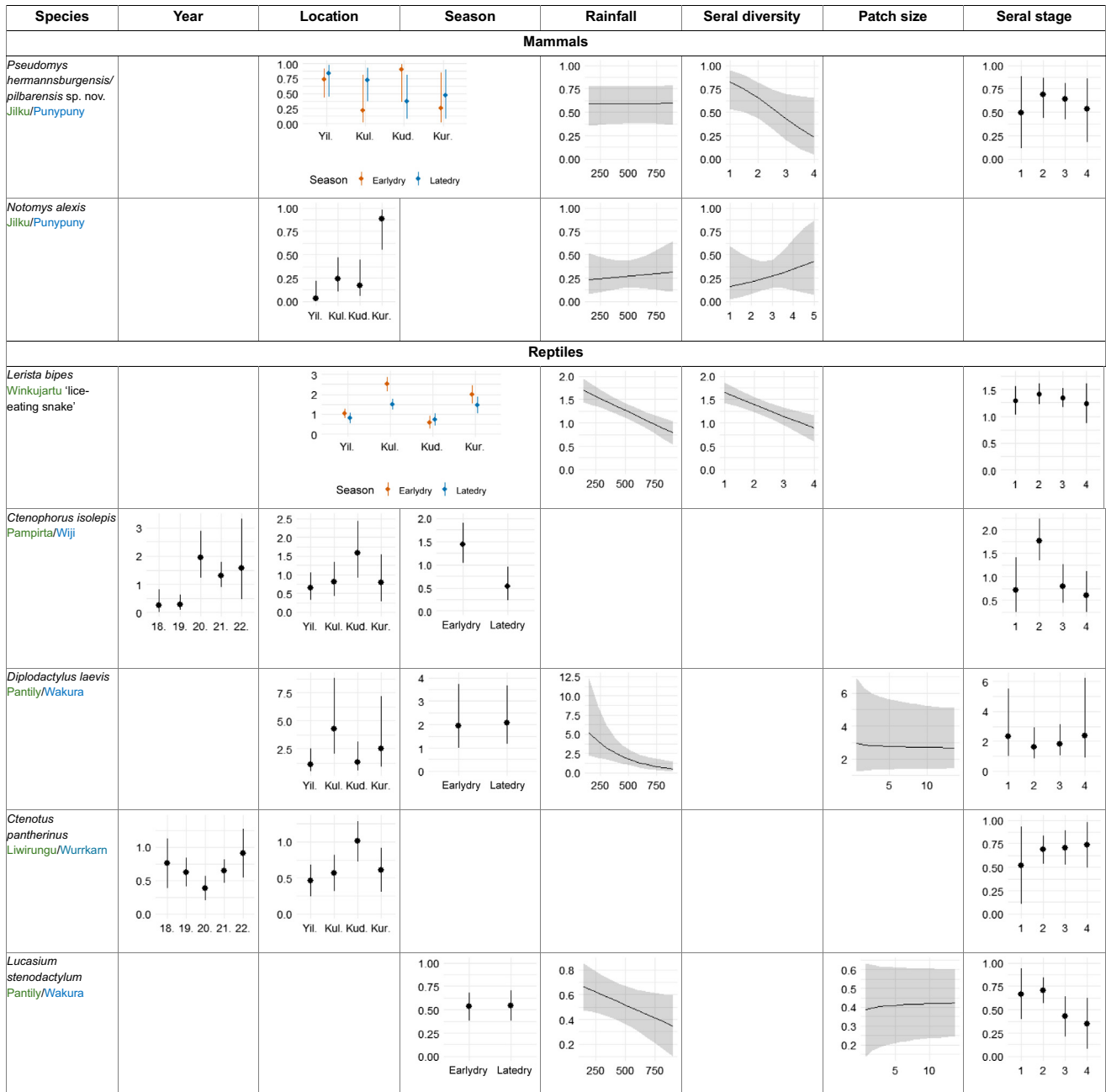


Fig. 6. Variables in the best models for the probability of occurrence for two mammal species, and in the daily capture rate for five reptile species. Values present estimated means and 95% confidence intervals. Graphs are included only for fixed effects in the best-ranked models. Statistics for fixed effects are in Table 3c.

tufts in dingo scats, in this area in the 1970s during surveys by Western Australian Government (McKenzie 1981). We detected *Larnkurr* on a camera trap image. Yilpi lies on the southern edge of the likely distribution of the northern subspecies, *T. v. arnhemensis* (northern brushtail possum), but records of *Larnkurr* from the northwest deserts are very few, and more clarity on the distributions of this subspecies plus those of the two other subspecies that could occur here (*T. v. vulpecula*, and *T. v. hypoleucas*) is needed (Woinarski

et al. 2014; TSSC 2021). Brushtail possums have declined in many areas, and the northern subspecies is listed as nationally Vulnerable, with declines attributed to changed fire regimes, and predation by cats (and foxes, at the southern edge of their range) (Woinarski *et al.* 2014; TSSC 2021).

We also detected kaluta, *Dasykaluta rosamondae*, at Kuduarra (on the Canning Stock Route) on Ngurrara Country. We caught this species four times. The kaluta has previously been recorded in the Pilbara region, only reaching the

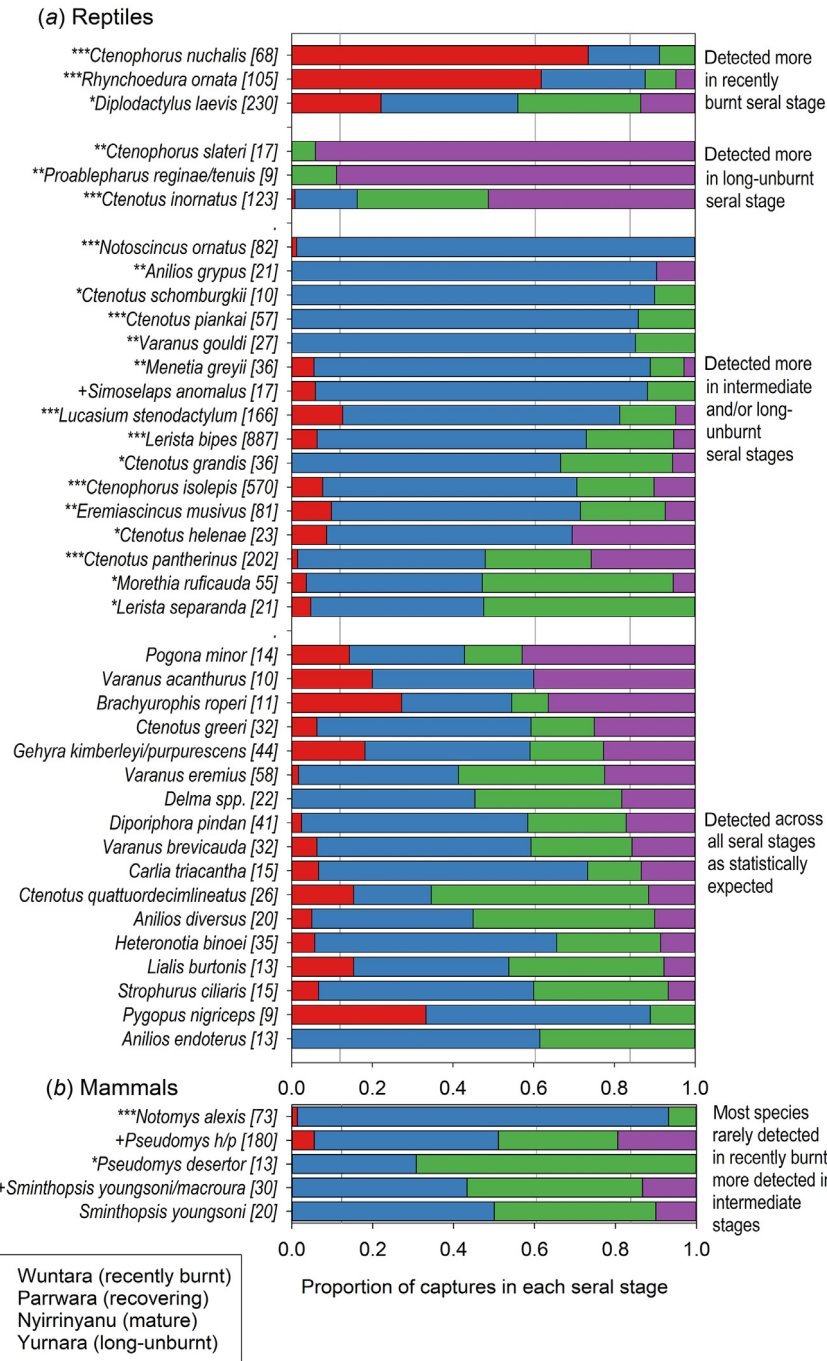


Fig. 7. The distribution of captures of (a) reptile and (b) mammal species across seral stages. In cases where the captures of species in a genus were low, and those species have similar ecologies, we grouped the captures in the genus to increase the sample size for analysis. *Pseudomys h/p* refers to *Pseudomys hermannsburgensis/pilbarensis* sp. nov.; *Delma* spp. refers to *D. borea*, *D. butleri*, *D. desmosa*, *D. nasuta*, *D. tincta*. Proportions of the overall sample of 81 site-surveys that belong to each of the four seral stages are shown by the vertical grey lines (0.12, 0.38, 0.24, 0.16 for recently burnt, recovering, mature and long-unburnt respectively). Total captures for each species are in square brackets after the species name. Species whose captures deviated significantly from expected using heterogeneity tests are marked as *** ($P < 0.0001$), ** ($P < 0.01$), * ($P < 0.05$), + ($P < 0.1$). The figure includes only those species with sufficient captures for statistical analysis; the full list of species caught, their captures, and the results of the heterogeneity tests, are in Table S1.

south-western edge of the Great Sandy Desert bioregion, so our observations extend the distribution known to scientists well into the Great Sandy Desert, by about 500 km. Similarly, we made many detections of the Dampierland plain slider, *Lerista separanda*, across all four locations on Karajarri and Ngurrara, with Kuduarrara being about 450 km away from existing records. The distribution of this species has been described by scientists as centred on the Dampierland bioregion (Chapple *et al.* 2019), but our records, and those of Doughty *et al.* (2011) from the Pilbara, suggest that the species may have a larger, desert-centred distribution that extends north into the Dampierland bioregion, and could mean that it no longer qualifies as a Priority 2-listed species in Western Australia (<https://bio.wa.gov.au/guide/conservation-status-definitions#p2>). The Karajarri name for *Lerista* spp is *Winkujarta*, lice-eating snake, because people used these legless skinks to clean lice out of their hair.

Seral stage and size

Mammal richness and capture rates declined immediately after fire, and increased over the next two seral stages, in a very similar pattern as the recovery of the spinifex layer itself (Fig. 2). Fire removes food resources (plant material and arthropods) for most mammals, and, by removing cover, exposes them to increased predation risk, including from introduced predators (McGregor *et al.* 2015; Leahy *et al.* 2016). The pattern in overall capture rates was driven by the three most-captured species, *Pseudomys hermannsburgensis/pilbarensis* sp. nov and *Notomys alexis*. These species have been reported to prefer recovering spinifex, 2–5 years after fire (Letnic and Dickman 2005; Bennison *et al.* 2018). However, less common mammal species seem to follow the same pattern, of being unlikely to occur in recently burnt habitats, and more likely to occur in older seral stages. *Pseudomys desertor* had the strongest preference for mature and long-unburnt seral stages, consistent with previous studies that have shown it to be more common in long-unburnt vegetation (Masters 1993; Letnic and Dickman 2005).

We found that 23 reptile species were mostly caught in certain seral stages. This was over half of the species with enough captures to test for seral preferences. This seral succession is likely to occur because some reptile species have preferences for microhabitat structure and floristics that are optimised at different stages of seral recovery (Masters 1996; Pianka 1996; Letnic *et al.* 2004; Pianka and Goodyear 2012). For example, shuttling heliotherm skinks need cover to protect them from the heat of the sun, and probably also from diurnal predators such as *Kirrkirri/Jarnun* (hawk, kite) (Pianka 1996); they are rarely present in recently burnt areas but re-establish as spinifex regrows (Letnic *et al.* 2004; Daly *et al.* 2008; Gordon *et al.* 2010). The clear sign of succession in our study is interesting, because fire intervals in the northern and western part of the Karajarri and Ngurrara Native Title areas can be as short as 4 or 5 years. Signs of succession are weaker in the

tropical savannas, potentially because the short fire intervals of 1–4 years leave little time for populations to establish as seral stages reach the preferred stage (Trainor and Woinarski 1994; Valentine and Schwarzkopf 2009). The strong successional patterns in our study despite rather short fire intervals suggest that most species are present across most areas, sometimes at undetectable levels of abundance, and populations can increase rapidly through *in situ* reproduction and/or emigration.

The seral preferences we observed have been reported previously (with minor variations) for *Ctenophorus nuchalis* and *Rhynchoedura ornata* (recently burnt); *Ctenophorus isolepis*, *Ctenotus helenae*, *Ctenotus grandis*, *Ctenotus pantherinus*, *Ctenotus piankai*, *Ctenotus schomburgkii*, *Lerista bipes*, *Notoscincus ornatus*, and *Menetia greyii* (recovering and/or mature seral stages) (Masters 1996; Letnic *et al.* 2004; Pianka and Goodyear 2012; Smith *et al.* 2013). The seral preferences we observed for several more species have not previously been reported (to our knowledge), but make sense given aspects of their ecology; *Proablepharus reginae*, *P. tenuis* and *Morethia ruficauda* are litter-dwellers. Leaf litter takes several years to redevelop after fire (Fig. 2), and we mostly caught these species in long-unburnt vegetation. *Ctenotus inornatus*, like many diurnal, shuttling heliotherm skinks, probably needs spinifex for shade and protection from predators including varanids, snakes, and cats (Greer 2024). This species was caught most often in long-unburnt vegetation. *Ctenophorus slateri* is usually found on small rocky outcrops (Greer 2024) that offer protection from fire, suggesting a degree of fire sensitivity, and we caught this species mostly in long-unburnt sites. *Diplodactylus laevis* is nocturnal, shelters in spider burrows, and is usually found in open habitats (Chapple *et al.* 2019; Greer 2024), and we found it preferred recently burnt areas. *Simoselaps anomalus* is a nocturnal sand-swimming snake that preys on small reptiles, especially *Lerista* spp. (Greer 2024), so its seral preference may follow that of its prey; both *S. anomalus* and *Lerista bipes* were mostly caught in recovering spinifex.

Although seral stage was important (in different ways) for both mammals and reptiles, we did not find strong evidence that the size of the seral patch was important. Patch size crept into one of the best-ranked models for reptile species richness, and for the capture rates of *Diplodactylus laevis* and *Lucasium stenodactylum*, but in all cases was not significant. One of the few desert studies that has looked at effects of patch size also found it to be uninfluential (Doherty *et al.* 2024). In contrast, mammal abundance and richness were higher in larger unburned areas of the Mojave Desert (Hulton VanTassel *et al.* 2015), and given the importance of scale in so many aspects of ecology, our result is surprising, and warrants further investigation (Mason and Lashley 2021).

Seral diversity

Given the seral preferences shown by many reptile species, it follows that areas containing mixed seral stages should

contain a greater variety of species assemblages, as has been shown for desert plants, which also show post-fire successional patterns (Wright and Clarke 2007; Greenwood *et al.* 2024). In addition, sites (such as our trapping sites) positioned within areas with more *surrounding* seral diversity could support higher species richness, for four reasons. First, the processes of colonisation into patches as they reach preferred stages should be faster with a greater proximity of source populations. Second, there could be higher rates of individuals 'leaking' into neighbouring, non-preferred seral stages. Third, some species are edge specialists; for example, the *Yalki/Wilka* (sand goanna, *Varanus gouldii*) has been shown to exploit the edges between seral stages (Bliege Bird *et al.* 2013). Fourth, seral diversity should increase the breadth of accessible resources within an animal's home range (e.g. annual grass seed in recently burnt areas adjacent to cover in mature and long-unburnt areas). Examples of species that may benefit from such fine-scale seral diversity include *Jitaru/Mirtuluju* (bilby, *Macrotis lagotis*) (Southgate *et al.* 2007); *Kuriti/Marrany* (dingo, *Canis familiaris dingo*) (Bliege Bird *et al.* 2018); and *Pikkarta/Jamarnti* (euro, *Osphranter robustus*) (Coddling *et al.* 2014).

However, our results on the relationships between seral diversity and biodiversity were mixed. We found no sign that seral diversity around a site, at small (500 m), medium (1 km) or large (5 km) scales, resulted in increased species richness or reptile captures at that site, and reptile capture rates decreased with seral diversity at the small scale. In contrast, seral diversity at the medium scale was positively associated with species richness in mammals. The difference between reptiles and mammals may be because mammals move further than most reptile species and are therefore more likely to cross seral boundaries (Dickman *et al.* 1995; Pianka 1996; Letnic 2001; Haythornthwaite and Dickman 2006). In addition, mammals may be more susceptible to threats, such as predation, that are themselves affected by seral diversity. For example, cats can be attracted to large, severe fires (that cause a single, extensive, early seral stage) while ignoring small, mild fires that result in a greater seral diversity (McGregor *et al.* 2014; Doherty *et al.* 2022). In addition, dingoes and goannas may increase activity in areas with a greater seral diversity. High dingo activity may cause a reduction in cat activity (Bliege Bird *et al.* 2013, 2018), and thus reduced predation pressure on small mammals because dingoes eat fewer small mammals than cats (Murphy *et al.* 2019).

Alternatively, areas with a high seral diversity may benefit biodiversity because they limit the spread, or affect the behaviour, of wildfire (Letnic and Dickman 2005; Pastro *et al.* 2011; Bliege Bird *et al.* 2012; D'Souza *et al.* 2013; Greenwood *et al.* 2022). Extensive, severe desert wildfires have been shown to cause reductions in species richness and abundance, most commonly in mammals (Letnic *et al.* 2004; Pastro *et al.* 2011; McDonald *et al.* 2016; Moore *et al.* 2024), but also in some reptiles, such as *Wurrkarn/Minijarti*, the great desert

skink (*Liopholis kintorei*), a culturally significant and nationally threatened species (Cadenhead *et al.* 2016; Indigenous Desert Alliance 2023). Key resources for ground-dwelling mammals and reptiles, such as woody debris, seeds, and invertebrates, are reduced or altered by fire (Wright and Clarke 2007; Pianka and Goodyear 2012). Species recovery after extensive, intense fires is slower across many animal groups including invertebrates (Langlands *et al.* 2012; Bastow 2020) and vertebrates (Shaw *et al.* 2021; Legge *et al.* 2022) potentially from depleted resources, but more likely from increased exposure to predation (Leahy *et al.* 2016; Moore *et al.* 2018; Doherty *et al.* 2022).

Mature and long-unburnt vegetation

Another feature of areas with high seral diversity is that they may contain and protect (by limiting or interacting with wildfire spread) areas of mature and long-unburnt habitat, which is important for a subset of species. To some extent, topographical features such as paleodrainages and rocky areas can create fire shadows that protect long-unburnt vegetation, but such areas are limited in the Great Sandy Desert, which is characterised by large unbroken expanses of sandplains and dunes. We found that up to eight reptile species were mostly, or almost only, found in *Nyirrinyanu*, mature vegetation, and *Yurnara*, long-unburnt vegetation, and several more species appeared to prefer these later seral stages (Fig. 6), but their overall captures were too low to be sure. All mammals except for *Notomys alexis* were more common in *Nyirrinyanu*, mature vegetation, and *Yurnara*, long-unburnt vegetation. *Notomys alexis* is able to evade predators even when ground cover is sparse with its fast hopping speed, and by sheltering in burrows (Spencer *et al.* 2014). However, even *N. alexis* was rarely caught in *Wuntara*, recently burnt vegetation.

The importance of long-unburnt vegetation for maintaining species assemblages in deserts recurs through an accumulating number of studies (Kelly *et al.* 2012; Langlands *et al.* 2012; Taylor *et al.* 2012; Nimmo *et al.* 2013; Farnsworth *et al.* 2014). Mature and long-unburnt vegetation appears essential for the persistence of some species, for example, *Pseudomys desertor*, *Dasycercus* spp., *Sminthopsis psammophila* (Masters 1993; Letnic and Dickman 2005; Riley *et al.* 2021), great desert skinks (Moore *et al.* 2015), and yulu, *Acanthophis pyrrhus* (McDonald *et al.* 2012).

Management implications

Our results help guide fire-management planning, not just for Karajarri and Ngurrara, but also other desert ranger groups. Australian deserts contain the largest network of Indigenous-managed land globally (O'Bryan *et al.* 2021). Our work can also guide other desert land managers including the Western Australian Department of Biodiversity, Conservation and Attractions. Our study encourages us to create a mix of seral stages across Country, so that there is good habitat for a range

of species. We should increase the area of Nyirrinyanu and Yurnara, mature and long-unburnt spinifex, because some species depend on this seral stage, yet it burns readily in wildfires. The average fire-return intervals across Karajarri and Ngurrara between 2000 and 2023 were 5.2 years and 7.1 years respectively (Fig. S1), but now we know that some species do best in spinifex that is at least 5 years old (5 years since fire). Our old people achieved older post-fire vegetation by creating a fine-scale patchwork of seral stages that tends to limit wildfire spread. We can aim to emulate this. The spatial scale of the patchwork may be less important for reptile species, but having multiple seral stages within a 1–5 km radius area (i.e. 6.3–31.4 km²) could benefit mammals. Since we began landscape-scale, coordinated fire management in 2019, we have reduced the mean size of fires by 52% (down to 2510 ha) and 40% (down to 21.6 km²) on Karajarri and Ngurrara respectively (Karajari Rangers, Ngurrara Rangers, unpubl. data). We can aim to drive those numbers lower.

Two-way collaboration

The Pirra Pirra Jungku/Pirra Warlu project has been a successful two-way collaboration between Traditional Custodians and scientists. The key feature underpinning this success has been the leadership of the Karajarri Rangers, and later also the Ngurrara Rangers. The rangers integrated the scientific method into their management, while maintaining authority over the project direction and conduct. The project aims, methods and results were discussed continuously by the Indigenous and non-Indigenous project partners, and co-presented to steering committees and Prescribed Body Corporates for ongoing endorsement. The rangers sought their own funding for the project, held all the required permits, and own the data. They have communicated the project at ranger meetings, scientific conferences and to government funders; hosted ranger exchanges so that trapping techniques could be learned peer-to-peer; and schooled the science partners in cultural dimensions of fire practice and looking after Country. The fire-history information and analyses have supported collaborative fire-management planning with neighbouring groups, pastoral stations, communities and elders, to discuss priorities, share resources, trouble-shoot issues such as weather events, access constraints and other pertinent operational steps. Overall, the project has built confidence in the ranger teams to lead and deliver large, long-term projects. It has also encouraged evolution in reporting, to include biodiversity outcomes of management as well as the more usual 'input' reporting metrics such as ranger time. As a result of its appeal, the Pirra Jungku/Pirra Warlu model is expanding to include other neighbouring desert ranger groups, including the Nyangumarta and Ngurrupa teams.

Supplementary material

Supplementary material is available [online](#).

References

- Allan G, Southgate R (2002) Fire regimes in the spinifex landscapes of Australia. In 'Flammable Australia: the fire regimes and biodiversity of a continent'. (Eds R Bradstock, J Williams, A Gill) pp. 145–176. (Cambridge University Press: Cambridge, UK)
- Bastow J (2020) The impacts of a wildfire in a semiarid grassland on soil nematode abundances over 4 years. *Biology and Fertility of Soils* **56**, 675–685. doi:10.1007/s00374-020-01441-4
- Bennison K, Godfree R, Dickman CR (2018) Synchronous boom–bust cycles in central Australian rodents and marsupials in response to rainfall and fire. *Journal of Mammalogy* **99**, 1137–1148. doi:10.1093/jmammal/ggy105
- Bent NJ, Chuguna JM, Lowe P, Richards E (2016) 'Two sisters.' (Magabala Books: Broome, WA, Australia)
- Bird DW, Bird RB, Parker CH (2005) Aboriginal burning regimes and hunting strategies in Australia's Western Desert. *Human Ecology* **33**, 443–464. doi:10.1007/s10745-005-5155-0
- Blackwood EMJ, Karajarri Rangers, Bayley S, Bijlani H, Fensham RJ, Lindsay M, Noakes E, Wemyss J, Legge S (2022) Pirra Jungku: comparison of traditional and contemporary fire practices on Karajarri Country, Western Australia. *Ecological Management & Restoration* **23**, 83–92. doi:10.1111/emr.12527
- Bliege Bird R, Codding BF, Kauhanen PG, Bird DW (2012) Aboriginal hunting buffers climate-driven fire-size variability in Australia's spinifex grasslands. *Proceedings of the National Academy of Sciences* **109**, 10287–10292. doi:10.1073/pnas.1204585109
- Bliege Bird R, Taylor N, Codding BF, Bird DW (2013) Niche construction and Dreaming logic: aboriginal patch mosaic burning and varanid lizards (*Varanus gouldii*) in Australia. *Proceedings of the Royal Society B: Biological Sciences* **280**, 20132297. doi:10.1098/rspb.2013.2297
- Bliege Bird R, Bird DW, Codding BF (2016) People, El Niño southern oscillation and fire in Australia: fire regimes and climate controls in hummock grasslands. *Philosophical Transactions of the Royal Society B: Biological Sciences* **371**, 20150343. doi:10.1098/rstb.2015.0343
- Bliege Bird R, Bird DW, Fernandez LE, Taylor N, Taylor W, Nimmo D (2018) Aboriginal burning promotes fine-scale pyrodiversity and native predators in Australia's Western Desert. *Biological Conservation* **219**, 110–118. doi:10.1016/j.biocon.2018.01.008
- Bowman DMJS (1998) The impact of Aboriginal landscape burning on the Australian biota. *New Phytologist* **140**, 385–410. doi:10.1111/j.1469-8137.1998.00289.x
- Bradstock RA (2010) A biogeographic model of fire regimes in Australia: current and future implications. *Global Ecology and Biogeography* **19**, 145–158. doi:10.1111/j.1466-8238.2009.00512.x
- Burbidge AA, McKenzie NL (1989) Patterns in the modern decline of Western Australia's vertebrate fauna: causes and conservation implications. *Biological Conservation* **50**, 143–198. doi:10.1016/0006-3207(89)90009-8
- Burrows ND, Christensen P (1991) A survey of Aboriginal fire patterns in the Western Desert of Australia. In 'Fire and the environment: ecological and cultural perspectives. Proceedings of an International Symposium, Knoxville, Tennessee, 20–24 March 1990. General Technical Report SE69'. (Eds SC Nodvin, TA Waldrop) pp. 297–310. (United States Department of Agriculture: Ashville, NC, USA)
- Burrows ND, Burbidge AA, Fuller PJ, Behn G (2006) Evidence of altered fire regimes in the Western Desert region of Australia. *Conservation Science Western Australia* **5**, 272–284.
- Cadenhead NCR, Kearney MR, Moore D, McAlpin S, Wintle BA (2016) Climate and fire scenario uncertainty dominate the evaluation of options for conserving the great desert skink. *Conservation Letters* **9**, 181–190. doi:10.1111/conl.12202
- Cameron L (2020) 'Healthy Country, Healthy People': Aboriginal Embodied Knowledge Systems in Human/Nature Interrelationships. *The International Journal of Ecopsychology (IJE)* **1**, 3.
- Chapple D, Tingley R, Mitchell N, Macdonald S, Keogh JS, Shea G, Bowles P, Cox N, Woinarski JCZ (2019) 'The action plan for Australian lizards and snakes 2017.' (CSIRO Publishing: Melbourne, Vic., Australia)
- Codding BF, Bliege Bird R, Kauhanen PG, Bird DW (2014) Conservation or co-evolution? Intermediate levels of aboriginal burning and hunting have positive effects on kangaroo populations in Western Australia. *Human Ecology* **42**, 659–669. doi:10.1007/s10745-014-9682-4

- Daly BG, Dickman CR, Crowther MS (2008) Causes of habitat divergence in two species of agamid lizards in arid central Australia. *Ecology* **89**, 65–76. doi:10.1890/06-1991.1
- Dickman CR, Predavec M, Downey FJ (1995) Long-range movements of small mammals in arid Australia: implications for land management. *Journal of Arid Environments* **31**, 441–452. doi:10.1016/S0140-1963(05)80127-2
- Doherty TS, Geary WL, Jolly CJ, Macdonald KJ, Miritis V, Watchorn DJ, Cherry MJ, Conner LM, González TM, Legge SM, Ritchie EG, Stawski C, Dickman CR (2022) Fire as a driver and mediator of predator–prey interactions. *Biological Reviews* **97**, 1539–1558. doi:10.1111/brv.12853
- Doherty TS, Bohórquez Fandiño DF, Watchorn DJ, Legge SM, Dickman CR (2024) Experimentally testing animal responses to prescribed fire size and severity. *Conservation Biology* **38**, e14231. doi:10.1111/cobi.14231
- Doughty P, Rolfe JK, Burbidge AH, Pearson DJ, Kendrick PG (2011) Herpetological assemblages of the Pilbara biogeographic region, Western Australia: ecological associations, biogeographic patterns and conservation. *Records of the Western Australian Museum, Supplement* **78**, 315–341. doi:10.18195/issn.0313-122x.78(2).2011.315-341
- D'Souza JB, Whittington A, Dickman CR, Leung LK-P (2013) Perfect storm: demographic responses of an irruptive desert mammal to prescribed burns following flooding rain. *Austral Ecology* **38**, 765–776. doi:10.1111/aec.12086
- Edwards GP, Allan GE, Brock C, Duguid A, Gabrys K, Vaarzon-Morel P (2008) Fire and its management in central Australia. *The Rangeland Journal* **30**, 109–121. doi:10.1071/RJ07037
- Farnsworth LM, Nimmo DG, Kelly LT, Bennett AF, Clarke MF (2014) Does pyrodiversity beget alpha, beta or gamma diversity? A case study using reptiles from semi-arid Australia. *Diversity and Distributions* **20**, 663–673. doi:10.1111/ddi.12181
- Gordon CE, Dickman CR, Thompson MB (2010) Partitioning of temporal activity among desert lizards in relation to prey availability and temperature. *Austral Ecology* **35**, 41–52. doi:10.1111/j.1442-9993.2009.02010.x
- Gould RA (1971) Uses and effects of fire among the Western Desert Aborigines of Australia. *Mankind* **8**, 14–24. doi:10.1111/j.1835-9310.1971.tb01436.x
- Greenwood L, Bliège Bird R, Nimmo D (2022) Indigenous burning shapes the structure of visible and invisible fire mosaics. *Landscape Ecology* **37**, 811–827. doi:10.1007/s10980-021-01373-w
- Greenwood L, Bliège Bird R, McGuire C, Jadai N, Price J, Skroblin A, van Leeuwen S, Nimmo D (2024) Indigenous pyrodiversity promotes plant diversity. *Biological Conservation* **291**, 110479. doi:10.1016/j.biocon.2024.110479
- Greer AE (2024) Encyclopedia of Australian Reptiles. Version: 1 June 2024. Available at <https://www.encyclopediaofaustralianreptiles.com/>
- Haythornthwaite AS, Dickman CR (2006) Long-distance movements by a small carnivorous marsupial: how *Sminthopsis youngsoni* (Marsupialia: Dasyuridae) uses habitat in an Australian sandridge desert. *Journal of Zoology* **270**, 543–549. doi:10.1111/j.1469-7998.2006.00186.x
- Hoffman KM, Davis EL, Wickham SB, Schang K, Johnson A, Larking T, Lauriault PN, Quynh Le N, Swerdfager E, Trant AJ (2021) Conservation of Earth's biodiversity is embedded in Indigenous fire stewardship. *Proceedings of the National Academy of Sciences* **118**, e2105073118. doi:10.1073/pnas.2105073118
- Hulton VanTassel HL, Barrows CW, Anderson KE (2015) Post-fire spatial heterogeneity alters ground-dwelling arthropod and small mammal community patterns in a desert landscape experiencing a novel disturbance regime. *Biological Conservation* **182**, 117–125. doi:10.1016/j.biocon.2014.11.046
- IDA (2023) Desert Habitat Method: accredited by Accounting for Nature. Indigenous Desert Alliance, 10 Deserts Project, Perth, WA, Australia.
- Indigenous Desert Alliance (2023) Looking after Tjakura, Tjalapa, Mulyamiji, Warrarna, Nampuu. A national recovery plan for the Great Desert Skink (*Liopholis kintorei*). Department of Climate Change, Energy, the Environment and Water, Canberra, ACT, Australia. Available at www.dcceew.gov.au/environment/biodiversity/threatened/recovery-plans
- Kelly LT, Nimmo DG, Spence-Bailey LM, Taylor RS, Watson SJ, Clarke MF, Bennett AF (2012) Managing fire mosaics for small mammal conservation: a landscape perspective. *Journal of Applied Ecology* **49**, 412–421. doi:10.1111/j.1365-2664.2012.02124.x
- Kimber R (1983) Black lightning: Aborigines and fire in central Australia and the Western Desert. *Archaeology in Oceania* **18**, 38–45. doi:10.1002/arco.1983.18.1.38
- KTLA (2014) Karajarri Healthy Country Plan 2013–2023: Palanapayana Tujjana Ngurra. Karajarri Traditional Lands Association, Bidyadanga, WA, Australia.
- Langlands PR, Brennan KEC, Ward B (2012) Is the reassembly of an arid spider assemblage following fire deterministic? *Austral Ecology* **37**, 429–439. doi:10.1111/j.1442-9993.2011.02299.x
- Leahy L, Legge SM, Tuft K, McGregor HW, Barmuta LA, Jones ME, Johnson CN (2016) Amplified predation after fire suppresses rodent populations in Australia's tropical savannas. *Wildlife Research* **42**, 705–716. doi:10.1071/WR15011
- Legge S, Rumpff L, Woinarski JCZ, Whiterod NS, Ward M, Southwell DG, Scheele BC, Nimmo DG, Lintermans M, Geyle HM, Garnett ST, Hayward-Brown B, Ensbey M, Ehmke G, Ah Yong ST, Blackmore CJ, Bower DS, Brizuela-Torres D, Burbidge AH, Burns PA, Butler G, Catullo R, Chapple DG, Dickman CR, Doyle KE, Ferris J, Fisher D, Gallagher R, Gillespie GR, Greenlees MJ, Hohnen R, Hoskin CJ, Hunter D, Jolly C, Kennard M, King A, Kuchinke D, Law B, Lawler I, Lawler S, Loyn R, Lunney D, Lyon J, MacHunter J, Mahony M, Mahony S, McCormack RB, Melville J, Menkhurst P, Michael D, Mitchell N, Mulder E, Newell D, Pearce L, Raadik TA, Rowley JLL, Sitters H, Spencer R, Valavi R, West M, Wilkinson DP, Zukowski S (2022) The conservation impacts of ecological disturbance: time-bound estimates of population loss and recovery for fauna affected by the 2019–2020 Australian megafires. *Global Ecology and Biogeography* **31**, 2085–2104. doi:10.1111/geb.13473
- Letnic M (2001) Long distance movements and the use of fire mosaics by small mammals in the Simpson Desert, central Australia. *Australian Mammalogy* **23**, 125–134. doi:10.1071/AM01125
- Letnic M, Dickman CR (2005) The responses of small mammals to patches regenerating after fire and rainfall in the Simpson Desert, central Australia. *Austral Ecology* **30**, 24–39. doi:10.1111/j.1442-9993.2004.01410.x
- Letnic M, Dickman CR, Tischler MK, Tamayo B, Beh C-L (2004) The responses of small mammals and lizards to post-fire succession and rainfall in arid Australia. *Journal of Arid Environments* **59**, 85–114. doi:10.1016/j.jaridenv.2004.01.014
- Lowe P (2002) 'Hunters and trackers.' (Rosenberg Publishing: Dural, NSW, Australia)
- Lundie-Jenkins G (1993) Ecology of the rufous hare-wallaby, *Lagorchestes hirsutus* Gould (Marsupialia: Macropodidae) in the Tanami Desert, Northern Territory. I Patterns of habitat use. *Wildlife Research* **20**, 457–475. doi:10.1071/WR9930457
- Martin HA (2006) Cenozoic climatic change and the development of the arid vegetation in Australia. *Journal of Arid Environments* **66**, 533–563. doi:10.1016/j.jaridenv.2006.01.009
- Mason DS, Lashley MA (2021) Spatial scale in prescribed fire regimes: an understudied aspect in conservation with examples from the southeastern United States. *Fire Ecology* **17**, 3. doi:10.1186/s42408-020-00087-9
- Masters P (1993) The effects of fire-driven succession and rainfall on small mammals in spinifex grassland at Uluru National Park, Northern Territory. *Wildlife Research* **20**, 803–813. doi:10.1071/WR9930803
- Masters P (1996) The effects of fire-driven succession on reptiles in spinifex grasslands at Uluru National Park, Northern Territory. *Wildlife Research* **23**, 39–47. doi:10.1071/WR9960039
- McDonald PJ, Luck GW, Pavey CR, Wassens S (2012) Importance of fire in influencing the occurrence of snakes in an upland region of arid Australia. *Austral Ecology* **37**, 855–864. doi:10.1111/j.1442-9993.2011.02348.x
- McDonald PJ, Stewart A, Schubert AT, Nano CEM, Dickman CR, Luck GW (2016) Fire and grass cover influence occupancy patterns of rare rodents and feral cats in a mountain refuge: implications for management. *Wildlife Research* **43**, 121–129. doi:10.1071/WR15220
- McGregor HW, Legge S, Jones ME, Johnson CN (2014) Landscape management of fire and grazing regimes alters the fine-scale habitat

- utilisation by feral cats. *PLoS ONE* 9, e109097. doi:10.1371/journal.pone.0109097
- McGregor H, Legge S, Jones ME, Johnson CN (2015) Feral cats are better killers in open habitats, revealed by animal-borne video. *PLoS ONE* 10, e0133915. doi:10.1371/journal.pone.0133915
- McKenzie N (1981) Wildlife of the Edgar Ranges area, south-west Kimberley, Western Australia. WA CALM, Perth, WA, Australia.
- McKenzie NL, Burbidge AA, Baynes A, Brereton RN, Dickman CR, Gordon G, Gibson LA, Menkhorst PW, Robinson AC, Williams MR, Woinarski JCZ (2007) Analysis of factors implicated in the recent decline of Australia's mammal fauna. *Journal of Biogeography* 34, 597–611. doi:10.1111/j.1365-2699.2006.01639.x
- Moore D, Kearney MR, Paltridge R, McAlpin S, Stow A (2015) Is fire a threatening process for *Liopholis kintorei*, a nationally listed threatened skink? *Wildlife Research* 42, 207–216. doi:10.1071/WR14227
- Moore D, Kearney MR, Paltridge R, McAlpin S, Stow A (2018) Feeling the pressure at home: predator activity at the burrow entrance of an endangered arid-zone skink. *Austral Ecology* 43, 102–109. doi:10.1111/aec.12547
- Moore H, Yawuru Country Managers, Bardi Jardi Oorany Ranger, Nyul Nyul Rangers, Nyikina Mangala Rangers, Gibson L, Dziminski M, Radford LJ, Corey B, Bettink K, Carpenter F, Mcphail R, Sonneman T, Greatwich B (2024) Where there's smoke there's cats: long unburnt habitat is crucial to mitigating the impacts of cats on the Ngarlgumirdi, greater bilby (*Macrotis lagotis*). *Wildlife Research* 51(5). doi:10.1071/WR23117
- Morton SR, Stafford Smith DM, Dickman CR, Dunkerley DL, Friedel MH, McAllister RRJ, Reid JRW, Roshier DA, Smith MA, Walsh FJ, Wardle GM, Watson IW, Westoby M (2011) A fresh framework for the ecology of arid Australia. *Journal of Arid Environments* 75, 313–329. doi:10.1016/j.jaridenv.2010.11.001
- Murphy SA, Paltridge R, Silcock J, Murphy R, Kutt AS, Read J (2018) Understanding and managing the threats to night parrots in south-western Queensland. *Emu - Austral Ornithology* 118, 135–145. doi:10.1080/01584197.2017.1388744
- Murphy BP, Woolley L-A, Geyle HM, Legge SM, Palmer R, Dickman CR, Augusteyn J, Brown SC, Comer S, Doherty TS, Eager C, Edwards G, Fordham DA, Harley D, McDonald PJ, McGregor H, Moseby KE, Myers C, Read J, Riley J, Stokeld D, Trewella GJ, Turpin JM, Woinarski JCZ (2019) Introduced cats (*Felis catus*) eating a continental fauna: the number of mammals killed in Australia. *Biological Conservation* 237, 28–40. doi:10.1016/j.biocon.2019.06.013
- Nimmo DG, Kelly LT, Spence-Bailey LM, Watson SJ, Taylor RS, Clarke MF, Bennett AF (2013) Fire mosaics and reptile conservation in a fire-prone region. *Conservation Biology* 27, 345–353. doi:10.1111/j.1523-1739.2012.01958.x
- O'Bryan CJ, Garnett ST, Fa JE, Leiper I, Rehbein JA, Fernández-Llamazares Á, Jackson MV, Jonas HD, Brondizio ES, Burgess ND, Robinson CJ, Zander KK, Molnár Z, Venter O, Watson JEM (2021) The importance of Indigenous Peoples' lands for the conservation of terrestrial mammals. *Conservation Biology* 35, 1002–1008. doi:10.1111/cobi.13620
- Pastro LA, Dickman CR, Letnic M (2011) Burning for biodiversity or burning biodiversity? Prescribed burn vs. wildfire impacts on plants, lizards, and mammals. *Ecological Applications* 21, 3238–3253. doi:10.1890/10-2351.1
- Pausas JG, Keeley JE (2009) A burning story: the role of fire in the history of life. *BioScience* 59, 593–601. doi:10.1525/bio.2009.59.7.10
- Perry JJ, Sinclair M, Wikmunea H, Wolmby S, Martin D, Martin B (2018) The divergence of traditional Aboriginal and contemporary fire management practices on Wik traditional lands, Cape York Peninsula, northern Australia. *Ecological Management & Restoration* 19, 24–31. doi:10.1111/emr.12301
- Pianka ER, Goodyear SE (2012) Lizard responses to wildfire in arid interior Australia: long-term experimental data and commonalities with other studies. *Austral Ecology* 37, 1–11. doi:10.1111/j.1442-9993.2010.02234.x
- Pianka ER (1996) Long-term changes in lizard assemblages in the Great Victoria Desert: dynamic habitat mosaics in response to wildfires. In 'Long-term studies of vertebrate communities'. (Eds ML Cody, JA Smallwood) pp. 191–215. (Academic: San Diego, CA, Australia)
- Pike J (2008) 'The art of fire.' (Backroom Press: Broome, WA, Australia)
- Putnis A, O'Leary P, Leach A, Ings E, See P (2021) Strong on Country: sustaining success in Indigenous land and sea management in Australia. Country Needs People Ltd, Canberra, ACT, Australia.
- R Core Team (2024) 'R: a language and environment for statistical computing [Computer software].' (R Foundation for Statistical Computing: Vienna, Austria). Available at <https://www.R-project.org/>
- Riley J, Turpin JM, Zeale MRK, Jayatilaka B, Jones G (2021) Diurnal sheltering preferences and associated conservation management for the endangered sandhill dunnart, *Sminthopsis psammophila*. *Journal of Mammalogy* 102, 588–602. doi:10.1093/jmammal/gyab024
- Roycroft E, Ford F, Ramm T, Schembri R, Breed WG, Burns PA, Rowe KC, Moritz C (2024) Speciation across biomes: rapid diversification with reproductive isolation in the Australian delicate mice. *Molecular Ecology* 33, e17301. doi:10.1111/mec.17301
- Ruscalleda-Alvarez J, Moro D, van Dongen R (2021) A multi-scale assessment of fire scar mapping in the Great Victoria Desert of Western Australia. *International Journal of Wildland Fire* 30, 886–898. doi:10.1071/WF21019
- Ruscalleda-Alvarez J, Cliff H, Catt G, Holmes J, Burrows N, Paltridge R, Russell-Smith J, Schubert A, See P, Legge S (2023) Right-way fire in Australia's spinifex deserts: an approach for measuring management success when fire activity varies substantially through space and time. *Journal of Environmental Management* 331, 117234. doi:10.1016/j.jenvman.2023.117234
- Shaw RE, James AI, Tuft K, Legge S, Cary GJ, Peakall R, Banks SC (2021) Unburnt habitat patches are critical for survival and in situ population recovery in a small mammal after fire. *Journal of Applied Ecology* 58, 1325–1335. doi:10.1111/1365-2664.13846
- Short J, Turner B (1994) A test of the vegetation mosaic hypothesis: a hypothesis to explain the decline and extinction of Australian mammals. *Conservation Biology* 8, 439–449. doi:10.1046/j.1523-1739.1994.08020439.x
- Smith AL, Michael Bull C, Driscoll DA (2013) Successional specialization in a reptile community cautions against widespread planned burning and complete fire suppression. *Journal of Applied Ecology* 50, 1178–1186. doi:10.1111/1365-2664.12119
- Southgate R, Paltridge R, Masters P, Carthew S (2007) Bilby distribution and fire: a test of alternative models of habitat suitability in the Tanami Desert, Australia. *Ecography* 30, 759–776. doi:10.1111/j.2007.0906-7590.04956.x
- Spencer EE, Crowther MS, Dickman CR (2014) Risky business: do native rodents use habitat and odor cues to manage predation risk in Australian deserts? *PLoS ONE* 9, e90566. doi:10.1371/journal.pone.0090566
- Taylor RS, Watson SJ, Nimmo DG, Kelly LT, Bennett AF, Clarke MF (2012) Landscape-scale effects of fire on bird assemblages: does pyrodiversity beget biodiversity? *Diversity and Distributions* 18, 519–529. doi:10.1111/j.1472-4642.2011.00842.x
- Thackway R, Cresswell ID (Eds) (1995) 'An Interim Biogeographic Regionalisation for Australia: a framework for establishing the national system of reserves, Version 4.0.' (Australian Nature Conservation Agency: Canberra, ACT, Australia)
- Trainer CR, Woinarski JCZ (1994) Responses of lizards to three experiments fires in the savanna forests of Kakadu National Park. *Wildlife Research* 21, 131–147. doi:10.1071/WR9940131
- TSSC (2021) Conservation advice *Trichosurus vulpecula arnhemensis* Northern Brushtail Possum. Department of Agriculture, Water and the Environment, Canberra, ACT, Australia. Available at https://www.environment.gov.au/cgi-bin/sprat/public/publicspecies.pl?taxon_id=83091
- Turner D, Lewis M, Ostendorf B (2011) Spatial indicators of fire risk in the arid and semi-arid zone of Australia. *Ecological Indicators* 11, 149–167. doi:10.1016/j.ecolind.2009.09.001
- Valentine LE, Schwarzkopf L (2009) Effects of weed-management burning on reptile assemblages in Australian tropical savannas. *Conservation Biology* 23, 103–113. doi:10.1111/j.1523-1739.2008.01074.x
- Weir JK, Stacey C, Youngetob K (2011) The benefits associated with caring for country: literature review. AITSIS, Canberra, ACT, Australia.
- Woinarski JCZ, Burbidge AA, Harrison PL (2014) 'The action plan for Australian mammals 2012.' (CSIRO Publishing: Melbourne, Vic., Australia)

- Woinarski JCZ, Burbidge AA, Harrison PL (2015) Ongoing unraveling of a continental fauna: decline and extinction of Australian mammals since European settlement. *Proceedings of the National Academy of Sciences* 112, 4531–4540. doi:10.1073/pnas.1417301112
- Wright BR, Clarke PJ (2007) Fire regime (recency, interval and season) changes the composition of spinifex (*Triodia* spp.)-dominated desert dunes. *Australian Journal of Botany* 55, 709–724. doi:10.1071/BT06240
- Wright BR, Laffineur B, Royé D, Armstrong G, Fensham RJ (2021) Rainfall-linked megafires as innate fire regime elements in arid Australian spinifex (*Triodia* spp.) grasslands. *Frontiers in Ecology and Evolution* 9, 666241. doi:10.3389/fevo.2021.666241
- Yanunijarra Aboriginal Corporation (2012) The Ngurrarawarnti Wulyu Martarnupurru 2012–2022 – Ngurrara Healthy Country Plan 2012–2022. Ngurrara Traditional Owners and Frank Weisenberger (KLC), Fitzroy Crossing, WA, Australia.

Data availability. Data are held by the Karajarri Rangers (Karajarri Traditional Lands Trust) and the Ngurrara Rangers (Yanunijarra Aboriginal Corporation).

Conflicts of interest. One of the authors (Sarah Legge) is an editor-in-chief with Wildlife Research and another (Hannah Cliff) was a guest editor for the special issue that includes this paper. To mitigate these potential conflicts of interest they were blinded from the review process.

Declaration of funding. This work was funded by the Australian Government's Indigenous Protected Area Program, Indigenous Ranger Program, and the National Environmental Science Program via the Threatened Species Recovery Hub; by the Western Australia State Natural Resource Management, Aboriginal Ranger Program, Lotteries West, and via inkind support from the Department of Biodiversity, Conservation and Attractions; by the Indigenous Desert Alliance/10 Deserts; and by the Australian Research Council (LP220200184).

Acknowledgements. We thank the Karajarri and Ngurrara Traditional Owners, and respectfully acknowledge their past and present elders, their custodianship of Country, and their support for this project. Many rangers, elders and coordinators have participated in these surveys, we thank them for their hard work and commitment. Led by Bayo Taylor and Jacko Shoveller, the Karajarri rangers included Eugene 'EB' Bumba, Frederick 'Gulu' Shoveller, Julian Nagomara, Kamahl Bangu, Marissa Munro, Paddy 'Moonie' Kittimbolt, Shannica 'Shani Bod' Boddington, Sheen Kitty, and Vanessa 'Beno' Kitty. Karajarri wish to especially mention Jess Bangu, a cultural leader and mentor for the ranger team and community, who passed away last year. Led by Chantelle Murray and Frankie McCarthy, the Ngurrara Rangers included Alexander Grace, Alfie Thirkall, Barnsey, Bevan Bent, Daren Smith, Darlene Kogolo, Edmond Smiler, Elton 'Toy' Smiler, Francis Thirkall, Isiah Jack, Jamin Bent, Justin Andrews, Lucas Brown, Lynette Rogers, Raylene 'Emily' Lenmardi, Regina Thirkall, Resheeda Thirkall, Stanley Thirkall, Sumayah Suprise, Thomas Nnarda, and Tracy Munda. Staff, interns and volunteers from Environs Kimberley, WA DBCA, the NESP Threatened Species Recovery Hub, the Indigenous Desert Alliance and the Kimberley Land Council gave time and expertise, and supported the project in diverse ways from refining fire mapping to digging pitfall traps to helping with communication products. Special shout out for Ed Blackwood, Annika Spiridis, Nico Rakotopare, and Jaana Dielenberg.

Author affiliations

^AResearch Institute for the Environment and Livelihoods, Charles Darwin University, Casuarina, NT 0810, Australia.

^BFenner School of Society and the Environment, The Australian National University, Acton, ACT 2602, Australia.

^CEnvirons Kimberley, 44 Blackman Street, Broome, WA 6725, Australia.

^DKarajarri Traditional Lands Association, Broome Lotteries House, Broome, WA 6725, Australia.

^EYanunijarra Aboriginal Corporation, Fitzroy Crossing, WA 6765, Australia.

^FIndigenous Desert Alliance, 587 Newcastle Street, West Perth, WA 6005, Australia.

^GDepartment of Biodiversity, Conservation and Attractions, Kimberley Region, Broome, WA 6725, Australia.

^HNorthern Australian Indigenous Land and Sea Management Alliance, 23 Ellengowan Drive, Brinkin, NT 0810, Australia.

^ISchool of Molecular & Life Sciences, Curtin University, GPO Box U1987, Perth, WA 6845, Australia.